

# Stable Isotope Biogeochemistry of Seabird Guano Fertilization: Results from Growth Chamber Studies with Maize (*Zea Mays*)

Paul Szpak<sup>1\*</sup>, Fred J. Longstaffe<sup>2</sup>, Jean-François Millaire<sup>1</sup>, Christine D. White<sup>1</sup>

<sup>1</sup> Department of Anthropology, The University of Western Ontario, London, Ontario, Canada, <sup>2</sup> Department of Earth Sciences, The University of Western Ontario, London, Ontario, Canada

## Abstract

**Background:** Stable isotope analysis is being utilized with increasing regularity to examine a wide range of issues (diet, habitat use, migration) in ecology, geology, archaeology, and related disciplines. A crucial component to these studies is a thorough understanding of the range and causes of baseline isotopic variation, which is relatively poorly understood for nitrogen ( $\delta^{15}\text{N}$ ). Animal excrement is known to impact plant  $\delta^{15}\text{N}$  values, but the effects of seabird guano have not been systematically studied from an agricultural or horticultural standpoint.

**Methodology/Principal Findings:** This paper presents isotopic ( $\delta^{13}\text{C}$  and  $\delta^{15}\text{N}$ ) and vital data for maize (*Zea mays*) fertilized with Peruvian seabird guano under controlled conditions. The level of  $^{15}\text{N}$  enrichment in fertilized plants is very large, with  $\delta^{15}\text{N}$  values ranging between 25.5 and 44.7‰ depending on the tissue and amount of fertilizer applied; comparatively, control plant  $\delta^{15}\text{N}$  values ranged between -0.3 and 5.7‰. Intraplant and temporal variability in  $\delta^{15}\text{N}$  values were large, particularly for the guano-fertilized plants, which can be attributed to changes in the availability of guano-derived N over time, and the reliance of stored vs. absorbed N. Plant  $\delta^{13}\text{C}$  values were not significantly impacted by guano fertilization. High concentrations of seabird guano inhibited maize germination and maize growth. Moreover, high levels of seabird guano greatly impacted the N metabolism of the plants, resulting in significantly higher tissue N content, particularly in the stalk.

**Conclusions/Significance:** The results presented in this study demonstrate the very large impact of seabird guano on maize  $\delta^{15}\text{N}$  values. The use of seabird guano as a fertilizer can thus be traced using stable isotope analysis in food chemistry applications (certification of organic inputs). Furthermore, the fertilization of maize with seabird guano creates an isotopic signature very similar to a high-trophic level marine resource, which must be considered when interpreting isotopic data from archaeological material.

**Citation:** Szpak P, Longstaffe FJ, Millaire J-F, White CD (2012) Stable Isotope Biogeochemistry of Seabird Guano Fertilization: Results from Growth Chamber Studies with Maize (*Zea Mays*). PLoS ONE 7(3): e33741. doi:10.1371/journal.pone.0033741

**Editor:** John P. Hart, New York State Museum, United States of America

**Received:** January 31, 2012; **Accepted:** February 21, 2012; **Published:** March 30, 2012

**Copyright:** © 2012 Szpak et al. This is an open-access article distributed under the terms of the Creative Commons Attribution License, which permits unrestricted use, distribution, and reproduction in any medium, provided the original author and source are credited.

**Funding:** The authors thank the Natural Sciences and Engineering Research Council of Canada, Canada Research Chairs Program and the Canada Foundation for Innovation for funding in support of this research. This is Laboratory for Stable Isotope Science Contribution #279. The funders had no role in study design, data collection and analysis, decision to publish, or preparation of the manuscript.

**Competing Interests:** The authors have declared that no competing interests exist.

\* E-mail: pszpak@uwo.ca

## Introduction

Seabird excrement (guano) was arguably the most economically significant organic fertilizer in the world prior to the twentieth century. The guano was mined from small, nearshore islands off the arid western coast of South America in the Peru-Humboldt upwelling region. The guano islands of Peru and Chile are typically composed of rocky cliffs essentially devoid of vascular plants, with a relatively small number of fauna (ants, spiders, scorpions, lizards) that are supported by allochthonous inputs from the guano birds (guano, carcasses, feathers, eggshells) [1]. Because the region receives virtually no precipitation, the guano accumulates in sedimentary layers. The once thick deposits of seabird guano (>50 m in some cases) were mined extensively during the guano boom of the 1800 s, and today the islands rarely have more than several years worth of droppings accumulated [1]. The trade in guano peaked during the middle of the nineteenth century, with

20 million tons being exported to Europe and North America between 1848 and 1875 [2]. The popularity of guano with European and North American farmers waned in the latter part of the nineteenth century for a number of reasons, including: increasing guano prices, irregular availability, unsuitability for particular crops (especially turnips), a dwindling supply, and the development of the chemical fertilizer industry [3]. In recent years, however, there has been a resurgence in its popularity (particularly in horticulture) as worldwide demand for organically grown produce has increased [4,5]. The importance of guano as a fertilizer prior to the nineteenth century is less well known, but is mentioned by Spanish chroniclers and in colonial administrative documents [6,7]. On this basis, some have suggested that it may have been of some importance in prehispanic agriculture [8,9].

From an ecological perspective, the importance of ornithogenic nitrogen to marine and terrestrial ecosystems has long been recognized [10–12]. A number of studies conducted in tropical,

temperate, subpolar, and polar regions have shown that seabird guano alters the concentration of soil nutrients (particularly  $\text{NH}_4^+$ ,  $\text{NO}_3^-$ ,  $\text{PO}_4^{3-}$ ,  $\text{K}^+$ ,  $\text{Mg}^{2+}$ ), plant tissue nutrients (N, P, K), and plant productivity [13–21]. Seabird guano may also affect the diversity of plant species present, though results from such studies are inconsistent [21]. Numerous factors other than the presence of guano may also affect the chemistry, physiology, and ecology of plants growing within or near seabird colonies. In field studies it is often difficult, or impossible, to rule out the effects of these factors, which include: physical disturbance caused by birds such as plant clipping or trampling [22,23], deposition of seabird carcasses, feathers and eggshells [24–26], and avian-aided seed dispersal [27].

Particularly large  $^{15}\text{N}$  enrichments in soils, plants, and animals (5–40%) have been recorded in and around seabird nesting sites, allowing for the relative contribution of avian-derived nutrients to be assessed (Table 1). Despite this large body of literature, there have been no investigations that examine the biogeochemical effects of seabird guano on the western coast South America, with

the majority of studies focusing on Oceania, Japan, California, and Antarctica [21]. Furthermore, no studies have addressed the isotopic biogeochemistry of seabird guano from an agricultural or horticultural standpoint. The purpose of this study, therefore, is to assess the isotopic and vital effects of Peruvian seabird guano fertilization on maize (*Zea mays*) under controlled conditions. In particular we examine the extent of the enrichment in plant  $^{15}\text{N}$  resulting from guano fertilization.

Plants are capable of utilizing several different soil N sources, both organic (amino acids) and inorganic ( $\text{NH}_4^+$ ,  $\text{NO}_3^-$ ,  $\text{N}_2$ ). From a biogeochemical perspective, the uptake, assimilation, and allocation/reallocation of N compounds are all significant. Uptake of  $\text{NO}_3^-$  in plant root cells occurs through at least three different  $\text{NO}_3^-$  transport systems [28]. Once inside the root,  $\text{NO}_3^-$  can be assimilated into organic N, or translocated to the shoot for assimilation by nitrate reductase (NR), nitrite reductase (NiR), and glutamine synthetase (GS) [29]. Little or no fractionation of  $^{15}\text{N}$  is reported to be associated with the uptake of  $\text{NO}_3^-$  [30–32];

**Table 1.** Summary of studies examining the effects of seabird guano on the isotopic composition ( $\delta^{15}\text{N}$ ) of plants and soils.

Location	Bird Species	Guano $\delta^{15}\text{N}$ (‰, AIR)	Plant $\delta^{15}\text{N}$ (‰, AIR)		Plant $\delta^{15}\text{N}$ (‰, AIR)		Reference
			Bird	Non-bird	Bird	Non-bird	
California	Mixed	–	36.4±2.7	5.3±0.9	33.6±2.0	7.2±1.2	[13]
California	Mixed	–	24.3 to 24.5	6.8 to 7.8	–	–	[173]
Falkland Islands and Antarctica	Mixed	10.9±1.9	–1.8 to 15.8	–7.9 to 7.6	14.0±0.3	0.0 to 9.7	[174]
Antarctica	Snow petrel	–	–3.1 to 25.8	–17.6 to –0.5	13.1 to 25.9	–13.4 to –1.0	[175]
Australia	Gentoo penguin	14.6	7.2 to 18.8	–9.8 to –0.7	–	–	[176]
North Africa	Various gulls	–	9.8 to 17.4	–	10.5 to 13.4	16.8 to 20.8	[51]
New Zealand	Westland petrel	–	–	–	14.1±0.3	–	[177]
New Zealand	Mixed	–	–3.9 to 9.1	–	–	–	[178]
New Zealand	Sooty shearwater	7.7	14.2±3.1	–6.1±1.7	–	–	[179]
Japan	Great cormorant	–	16.4 to 16.9	–2.5±0.6	10.6 to 16.0	0.4±0.3	[180]
Japan	Great cormorant	13.2±1.3	10.0 to 14.7	–2.3 to 6.8	–	–	[181]
Sweden	Great cormorant	–	13.6±1.7	1.7	–	–	[182]
New Zealand	Mixed	–	4.6 to 6.7	14.4 to 15.9	10.2±1.0	16.2±0.3	[183]
Fiji	Mixed	39.1 to 50.1 <sup>a</sup>	13.6 to 36.7	–1.3 to 0.8	15.1 to 31.6 <sup>a</sup>	–4.1 to –1.3 <sup>b</sup>	[184]
Fiji	Mixed	14.9 to 23.3 <sup>b</sup>	–	–	13.5 to 33.0 <sup>b</sup>	–	[184]
Japan	Black-tailed gull	10.2 to 10.5	3.9 to 14.6	–4.1 to –2.1	10.1 to 43.3 <sup>a</sup>	–4.3 to –2.9 <sup>b</sup>	[185]
Japan	Black-tailed gull	–	–	–	–0.2 to 33.7 <sup>b</sup>	–	[185]
Japan	Black-tailed gull	9.1 to 12.8	–	–	18.5 to 44.1 <sup>a</sup>	–	[53]
Japan	Black-tailed gull	–	–	–	–4.1 to 42.2 <sup>b</sup>	–	[53]
Japan and Antarctica	Penguin and gull	8.0 to 9.4	13.6 to 38.1	–	–	–	[186]
Antarctica	Penguin	7.4	–	–	32.1	–	[75]
Japan	Mixed	–	–	–	9.1 to 37.9	–4.6 to 8.6	[54]
New Zealand	Rockhopper penguin	7.0±0.4	–	–	23.8±3.3	–0.5±0.2	[187]
Australia	Mixed	9.9	9.5±2.2	7.0±2.6	–	–	[55]
California	Mixed	–	27.2 to 27.3	8.3 to 9.5	28.3±5.4	–	[188]
Pribilof Islands	Mixed	12.5	22.0	11.3	–	–	[162]
California	Mixed	–	–	–	35.6±2.6	7.5±0.3	[56]
Pacific (Palmyra Atoll)	Mixed	13.9	14.0±1.4	9.3±0.9	16.2±0.3	11.0±0.7	[189]
Antarctica	Penguin	20.9±4.2	–	–	10.4±3.1	–	[190]

<sup>a</sup> $\text{NH}_4^+$ .

<sup>b</sup> $\text{NO}_3^-$ .

doi:10.1371/journal.pone.0033741.t001

fractionation of  $^{15}\text{N}$  does not appear to vary with respect to source  $[\text{NO}_3^-]$  [33–35]. Some variability in fractionation is associated with NR activity, and it has been difficult in some cases to differentiate between isotopic fractionation associated with N uptake and assimilation, respectively [36]. Ledgard et al. [37] report the fractionation for the entire process to be  $-15\%$ , while a range of 0 to  $-19\%$  is reported by Robinson [38].

$\text{NH}_4^+$  is taken up by plants via high or low affinity transporters depending on extracellular  $[\text{NH}_4^+]$  [39].  $\text{NH}_4^+$  is assimilated into organic N only in the roots via GS and most estimated  $\Delta^{15}\text{N}$  values for  $\text{NH}_4^+$  uptake and assimilation fall between  $-5$  and  $-20\%$  [38,40]. Unlike  $\text{NO}_3^-$ , however, there are substantial differences in  $\Delta^{15}\text{N}$  with source  $[\text{NH}_4^+]$ . For example, in two different rice cultivars, Yoneyama et al. [33] found  $\Delta^{15}\text{N}$  for  $\text{NH}_4^+$  uptake to be  $-6.1$  to  $-12\%$  at low source  $[\text{NH}_4^+]$ , and  $-13.4$  to  $-28.9\%$  at high source  $[\text{NH}_4^+]$ .

## Materials and Methods

### Materials

All plants were grown in a walk-in growth chamber at the Biotron Centre for Experimental Climate Change Research at the University of Western Ontario. The substrate utilized for all treatments was Pro-mix<sup>®</sup> for containers (75–85% sphagnum moss, 15–25% perlite and limestone). Peruvian seabird guano (Guano Company International, Cleveland, Ohio, United States) was obtained from an organic gardening outlet. The nitrogen content of the guano was reported to be 10% and determined to be  $11.2 \pm 0.2\%$  based on five analyses of dried, powdered guano as described for plant samples below. The ‘Early Sunglow’ maize cultivar was used (*Zea mays* cv. Early Sunglow, Lot E1, 2010, Ferry Morse, Fulton, Kentucky, United States) for all experiments because it is a relatively small variety of maize that accommodated physical restrictions on plant height imposed by the growth chamber.

### Growth Chamber Conditions

Growth chamber temperature was 25/18°C (day/night), with a photoperiod of 13 h provided by 185 W fluorescent bulbs. Relative humidity was set at 80% for the first four daylight hours, and 60% for the remainder of the day. These conditions were monitored electronically, and did not deviate from these parameters for the duration of the experiment.

### Maize Germination Experiment

Guano (well-mixed with soil) was applied to 1.2 L plastic containers (1.0 L of soil) in the following amounts: 0 g, 1.0 g, 2.5 g, 5.0 g, 7.5 g, 10.0 g and 15.0 g. Six replicates of each treatment were prepared. One hour after addition of the guano, maize seeds were planted  $\sim 2.5$  cm below the surface in the containers. Emergence and growth of the plants were recorded every 2–3 days for 35 days.

### Maize Fertilization Experiment

Fifteen maize seeds were planted  $\sim 2.5$  cm below the surface in 1.2 L plastic containers (1.0 L of soil). At this time, guano was mixed with soil in free-draining (perforated at the base) 18.9 L plastic buckets containing 16 L of soil in the following amounts: 0 (C0), 80 g (G1, 5 g guano/L), 160 g (G2, 10 g guano/L). Five replicates of each treatment were prepared. Maize is typically fertilized prior to planting, and sometimes again approximately three weeks after emergence, although this second application is uncommon [41]. To avoid complications associated with additional fertilizer applications, only one fertilizer application was

employed. After germination (7 days after sowing) maize plants were moved into the 18.9 L plastic buckets. Plants were watered every 2–3 days and the height and general growth of the plants was monitored. Distal leaf samples ( $\sim 3$  cm  $\times$  6 cm) were taken at 30 and 75 days after planting (d). Plants at 30 d were characterized by only vegetative growth, while plants sampled at 75 d had begun reproductive growth (tassels fully emerged, silks beginning to appear). Anthers were sampled at 75 d. At completion of the experiment (115 d), the following tissues were sampled: leaves, grains, roots, and stalks. All buckets were relocated randomly within the growth chamber five times (30, 45, 60, 75, 100 d) during the course of the experiments to account for any micro-variations in light, temperature or humidity, although such changes were not expected.

### Stable Isotope Analysis

All plant materials were stored at  $-25^\circ\text{C}$  following sampling until needed for analysis. Samples were then dried at  $90^\circ\text{C}$  under normal atmosphere for 72 hours, ground using a Wig-L-Bug (Crescent, Lyons, Illinois, United States) and the resulting powders stored at room temperature in sealed glass vials. Isotopic compositions ( $\delta^{13}\text{C}$  and  $\delta^{15}\text{N}$  values determined separately) and relative percentages of carbon and nitrogen were determined using a Delta V isotope ratio mass spectrometer (Thermo Scientific, Bremen, Germany) coupled to an elemental analyzer (Costech Analytical Technologies, Valencia, California, United States). For the analysis of  $\delta^{15}\text{N}$ , excess  $\text{CO}_2$  was removed using a Carbo-Sorb trap (Elemental Microanalysis, Okehampton, Devon, United Kingdom). Sample reproducibility was  $\pm 0.09\%$  for  $\delta^{13}\text{C}$  and  $\pm 0.90\%$  for %C (6 replicates), and  $\pm 0.12\%$  for  $\delta^{15}\text{N}$  and  $\pm 0.10\%$  for %N (24 replicates). A  $\delta^{15}\text{N}$  value of  $20.31 \pm 0.18\%$  was obtained for 37 analyses of IAEA-N2, which compared well with its accepted value of 20.30%. A  $\delta^{13}\text{C}$  value of  $-29.87 \pm 0.29\%$  was obtained for 11 analyses of NBS-22, which compared well with its accepted value of  $-30.00\%$ .

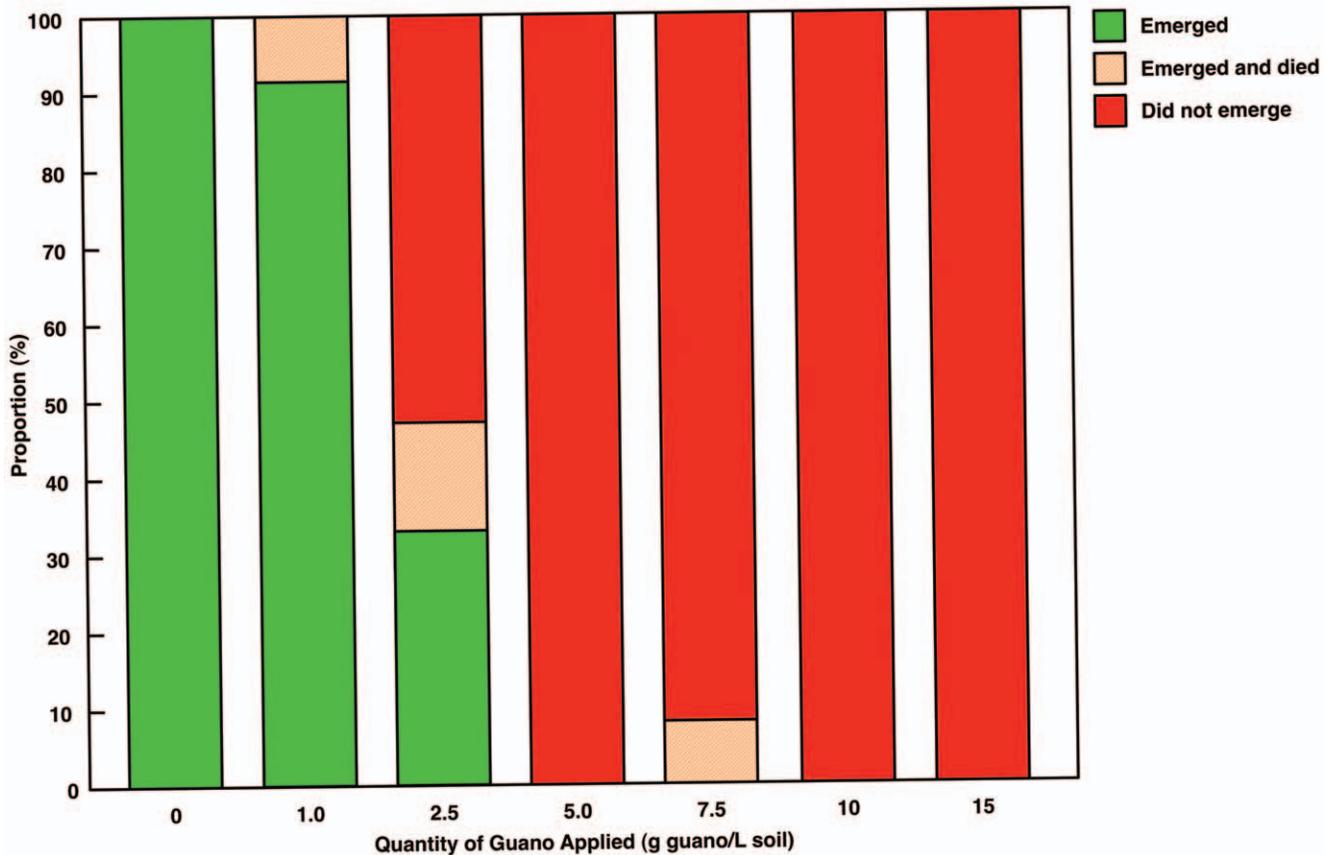
### Statistical Analyses

Comparisons between treatments and between organs were completed using one-way analysis of variance (ANOVA). Levene’s test was used to assess homogeneity of variance; if variance was homoscedastic, a *post hoc* Tukey’s honestly significant difference (HSD) test was applied and if variance was not homoscedastic, a *post hoc* Dunnett’s T3 test was applied. All statistical analyses were conducted at a significance level of 5% ( $p < 0.05$ ). All statistical analyses were performed in SPSS 16 for Windows.

## Results and Discussion

### Maize Germination and Seedling Establishment

All unfertilized plants germinated and commenced normal growth (Figure 1). There was a clear trend towards the inhibition of germination and seedling emergence with increasing rate of guano applied (Figure 1). It is apparent that the presence of seabird guano in the soil has the potential to inhibit germination and that this effect is concentration dependent. Ishida [42] found lower germination rates in oak and pine trees within, compared to outside of, cormorant colonies but did not offer a detailed explanation for this pattern. Mulder and Keall [43] also found that seabird guano negatively affected seed germination and seedling survival. Germination inhibition with increasing concentrations of guano probably results from a number of factors, including reduced soil pH and the presence of a high concentration of soluble salts, both of which are characteristic of ornithogenic soils [20]. Very high concentrations of  $\text{NO}_3^-$  and especially  $\text{NH}_4^+$  are



**Figure 1. Relative percentages of seedlings that germinated and emerged with differing amounts of seabird guano applied.**  
doi:10.1371/journal.pone.0033741.g001

also characteristic of ornithogenic soils and these characteristics can inhibit maize germination [44], with the early stages of growth being the most detrimental for plants under  $\text{NH}_4^+$  stress [45,46].

### Vital Effects of Guano Fertilization

Plant growth was strongly inhibited in the heavy guano treatment (G2). Maximum plant heights were significantly lower in G2 compared to C0 ( $p=0.02$ ) and G1 ( $p=0.008$ ) (Figure 2). While the G1 plants did not attain greater maximum heights than the C0 plants ( $p=0.83$ ), they yielded significantly more grain ( $p=0.004$ ). The G2 plants yielded less grain than the G1 plants ( $p=0.03$ ) and more grain than the C0 plants, although this difference was not statistically significant ( $p=0.42$ ) (Figure 2).

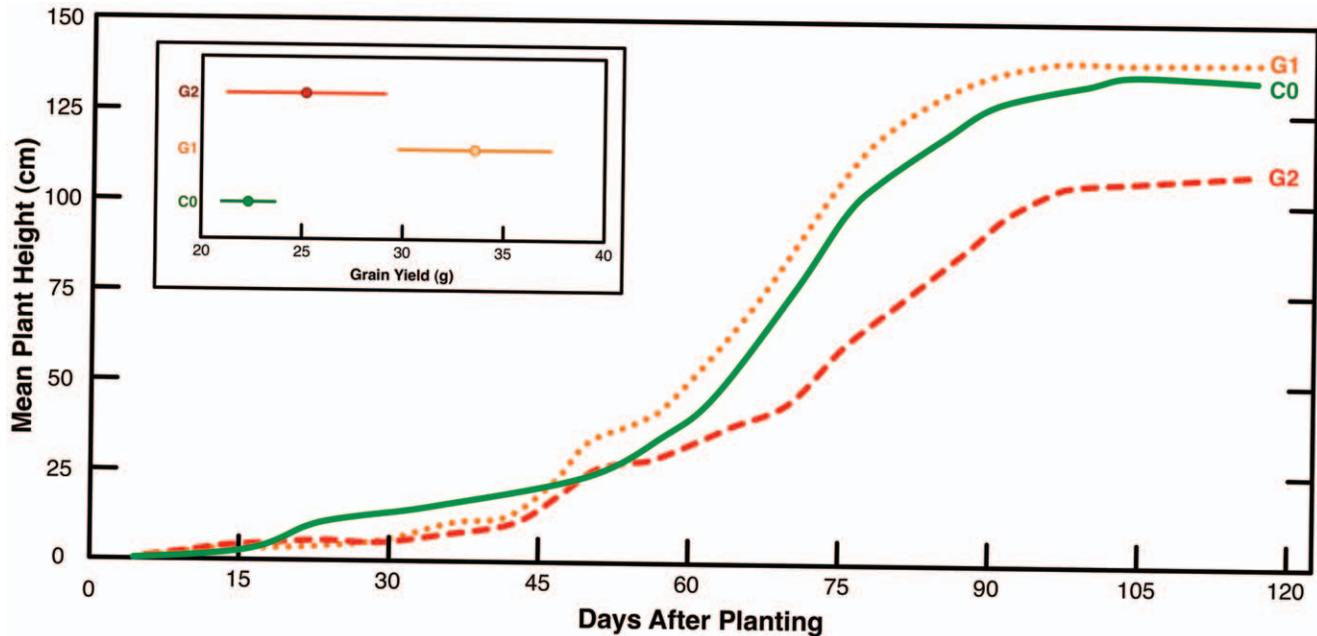
In this study, we observed a positive influence of guano on maize growth at moderate concentrations (G1), but a negative influence at high concentrations (G2). A number of studies have found that plant abundance and/or species richness tends to be lower within seabird colonies, but is often higher in areas in relatively close proximity to the colonies [14,22,47,48].

Very high levels of soil P can have deleterious effects on plant growth [49]. Ornithogenic soils may contain fifty times more available phosphorus than normal, but the P salts in bird excrement tend to be immobile in soil because of their limited solubility, making them generally unavailable for uptake by plants [4,50,51]. It is thus unlikely that the reduced growth observed in the G2 plants is the result of P toxicity. The most likely cause for the reduced growth of the G2 plants is  $\text{NH}_4^+$  toxicity.

Very high  $[\text{NH}_4^+]$  is a ubiquitous trait of ornithogenic soils [52–56]. High soil  $\text{NH}_4^+$  can negatively impact plants in several ways:

(1) soil acidification, particularly of the rhizosphere [57], ‘scorching’ of root hairs [46]; (2) accumulation of free  $\text{NH}_4^+$  in plant tissues, which has the capacity to uncouple plastid energy gradients [46]; (3) assimilation of  $\text{NH}_4^+$  in the roots and associated translocation of carbon skeletons from the shoot, which is metabolically expensive and places ‘carbon stress’ on roots [58]; (4) suppression of the expression of certain proteins (aquaporins), which can have detrimental effects on the uptake of water [59]; and (5) the influx and efflux of  $\text{NH}_4^+$  through root cells, which is associated with a very high metabolic cost when source  $[\text{NH}_4^+]$  is high [60].

Both the G1 and G2 plants exhibited significantly reduced growth compared to the control plants for the first 45 days of the experiment (Figure 2;  $p=0.01$ ), but this trend did not continue as the G1 plants produced the greatest yields, and had similar maximum heights to the control plants. This is likely the result of initially very high soil  $[\text{NH}_4^+]$ , which negatively impacted the growth of the fertilized plants, followed by increased soil  $\text{NO}_3^-$  resulting from nitrification of guano-derived  $\text{NH}_4^+$ . When plants largely supplied with  $\text{NH}_4^+$  as an N source are supplemented with  $\text{NO}_3^-$ ,  $\text{NH}_4^+$  uptake is suppressed and plants are able to resume normal growth [58]. The fact that the G2 plants still produced grain even though they were characterized by reduced heights and less above-ground biomass than either the control or G1 plants suggests that there was some acclimatization of these plants to the high  $[\text{NH}_4^+]$ , and/or nitrification was substantially delayed and  $[\text{NH}_4^+]$  remained high in the soil for a much longer period of time. Schortemeyer et al. [46] observed a similar result in maize plants grown with  $\text{NH}_4^+$  as the sole N source.



**Figure 2. Maximum heights of maize plants throughout experiment.** Harvest occurred at 115 d. Inset: grain yield for each experiment. doi:10.1371/journal.pone.0033741.g002

The effects of guano on plants are difficult to generalize. There is considerable variability at the community level and also within a community in accordance with plant physiology (nutrient demands, salt tolerance) at the species level [47,52]. Even within maize there are differences in  $\text{NH}_4^+$  tolerance, with some varieties being able to survive higher concentrations than others [46]. Therefore, it cannot be assumed that the results of this study are directly applicable to all maize varieties.

### Nitrogen Isotope Composition of Seabird Guano

Most inorganic N fertilizers have  $\delta^{15}\text{N}$  values close to 0‰, with organic fertilizers generally having highly variable but positive  $\delta^{15}\text{N}$  values (Table 2). The  $\delta^{15}\text{N}$  value of the seabird guano used in this experiment was  $26.7 \pm 0.6\text{‰}$  (5 analyses), which is much higher than any other organic fertilizer analyzed to date. This is the product of avian nitrogen metabolism and excretion, which is quite different than in mammals, combined with the high trophic position of the guano-producing birds. Guano contains 9–21% nitrogen, which is composed primarily of uric acid (~80%), with smaller amounts of protein (~10%), ammonia (~7%), and nitrate (~0.5%) [23,61–67]. In addition, guano contains ~4% phosphorous (~50% of which is  $\text{PO}_4^{3-}$ ) and 2% potassium [62,67,68].

A simplified pathway for guano nitrogen, with associated nitrogen-isotope fractionation factors, is shown in Figure 3. The principal producers of guano on the western coast of South America are the Peruvian booby (*Sula variegata*), brown pelican (*Pelecanus occidentalis thagus*), and guanay cormorant (*Phalacrocorax bougainvillii*) [1,12]. These birds, and similar species, feed at high trophic levels, and typically have tissue  $\delta^{15}\text{N}$  values in the range of 17 to 20‰ [69–71], suggesting a  $\delta^{15}\text{N}_{\text{diet}}$  of 14 to 18‰ assuming a diet–tissue fractionation of 3–4‰ for  $\delta^{15}\text{N}$  [72]. Thus, the high trophic level of the birds only partially explains the very high  $\delta^{15}\text{N}_{\text{bulk guano}}$  of 26.7‰.

After deposition in the soil, the uric acid in guano is rapidly mineralized to  $\text{NH}_4^+$ , and this process occurs much more rapidly in the presence of water [68,73,74]. Based on results presented by Mizutani and Wada [65], uric acid quickly decomposed (75% in

ten days) in soil, but the  $\delta^{15}\text{N}$  value of the remaining uric acid was unchanged. A very large isotopic fractionation (–40 to –60‰) occurs during  $\text{NH}_3$  volatilization, leaving the remaining soil  $\text{NH}_4^+$  highly enriched in  $^{15}\text{N}$  [38,75]. Ammonia volatilization is largely responsible for the high  $\delta^{15}\text{N}$  values in ornithogenic soils and in some cases, seabird guano (Table 1). The relatively high  $\delta^{15}\text{N}$  value of the guano utilized in this study suggests that some of the  $\text{NH}_4^+$  in the guano had been subject to volatilization prior to deposition in the soil during the experiment; similar observations have been made concerning other avian manures [76].

### $^{15}\text{N}$ Enrichment in Guano Fertilized Plants

Plant isotopic compositions are summarized in Table 3; raw data are presented in Table S1. Plant organs of fertilized plants (G1, G2) sampled at 115 d were significantly enriched in  $^{15}\text{N}$  compared to control plants in every case (Tables 3, 4; Figure 4). Also, the  $\delta^{15}\text{N}$  values of plant tissues were significantly higher for heavily fertilized (G2) versus more lightly fertilized (G1) plants (Tables 3, 4). The difference in mean  $\delta^{15}\text{N}$  values between the G1 and G2 plant organs was fairly consistent: 6.2‰ for stalks and roots, 6.4‰ for leaves (at 115 d), 7.6‰ for grain, and 7.8‰ for anthers.

A growing body of literature has emerged in recent years demonstrating that organic fertilizers, specifically those derived from animal waste, can cause large  $^{15}\text{N}$  enrichments of plant tissues (Table 5). The  $\delta^{15}\text{N}$  values reported here for plants grown in guano-fertilized soils are significantly higher than any published  $\delta^{15}\text{N}$  values for plants grown on other organic fertilizers to date (Table 5), but comparable to  $\delta^{15}\text{N}$  values for plants growing in ornithogenic soils (Table 2). The higher  $\delta^{15}\text{N}$  values in the G1 and G2 compared to the C0 plants is the result of the uptake of  $^{15}\text{N}$ -enriched guano-derived nitrogen. Moreover, the significantly higher tissue  $\delta^{15}\text{N}$  values in the G2 compared to G1 plants reflects, at least in part, the greater availability of guano-derived nitrogen throughout the course of the experiment. This does not imply that guano-derived N was absent in the G1 treatment towards the end of the experiment, but it is possible that N

**Table 2.**  $\delta^{15}\text{N}$  values of organic and inorganic fertilizers.

Type	Fertilizer	Fertilizer $\delta^{15}\text{N}$ (‰, AIR)	Reference
Organic	Blood	6.0±1.3	[191]
	Bonemeal	4.9±0.3	[191]
	Cattle manure	5.0±0.8	[192]
	Cattle manure	2.9±0.5	[193]
	Cattle manure	4.5	[133]
	Cattle manure	3.1±0.2	[85]
	Chicken manure	6.2±1.9	[191]
	Fishmeal	7.1±3.6	[191]
	Hoof and horn	6.4±0.2	[191]
	Livestock manure	8.8±4.4	[191]
	Livestock manure	8.7±0.2	[132]
	Pig manure	13.9	[102]
	Pig manure	16.9	[194]
	Pig manure	11.3	[133]
	Pig manure	6.5	[133]
	Pig manure	16.4	[195]
	Poultry manure	8.6±0.3	[132]
	Poultry manure	2.7	[133]
	Seabird guano	26.7±0.6	This study
Seaweed	2.5±1.5	[191]	
Various composts	17.4±1.2	[196]	
Inorganic	(NH <sub>4</sub> ) <sub>2</sub> H <sub>2</sub> PO <sub>4</sub>	-0.6±0.4	[191]
	(NH <sub>4</sub> ) <sub>2</sub> SO <sub>4</sub>	1.7±3.4	[191]
	(NH <sub>4</sub> ) <sub>2</sub> SO <sub>4</sub>	-1.6	[133]
	(NH <sub>4</sub> ) <sub>2</sub> SO <sub>4</sub>	-2.6	[197]
	KNO <sub>3</sub>	-1.2±0.3	[191]
	NH <sub>4</sub> NO <sub>3</sub>	-1.3	[128]
	NH <sub>4</sub> NO <sub>3</sub>	-0.6±1.7	[191]
	NH <sub>4</sub> NO <sub>3</sub>	-1.7	[133]
	Urea	-2.4±2.1	[191]
	Urea	-1.7	[133]
	Urea	-0.7	[195]

doi:10.1371/journal.pone.0033741.t002

immobilization had overtaken N mineralization, reducing the amount of guano-derived N available to the plants.

### Elemental Concentration in Plant Parts

There were significant differences in N content between fertilized and control plants, with fertilized plants tending to have significantly higher N (Tables 3, 4). There were no significant differences in C content between control and fertilized plants for all organs, with the exception of the stalks, which had significantly lower %C in the fertilized plants compared to the control, and in G2 compared to G1 plants.

In general, the differences in C and N content between fertilized and unfertilized plants can be attributed to the accumulation of proteins, particularly those related to the GS-GOGAT pathway, that assimilate NH<sub>4</sub><sup>+</sup> and amino acids. Free amino acids tend to accumulate unabated in plant tissues with increasing supply of N irrespective of source, although different amino acids may accumulate at different rates depending on plant species and N

source [45,77–80]. Moreover, many studies have noted an increase in proteins, such as GS, in plant tissue in accordance with increasing NH<sub>4</sub><sup>+</sup> supply [80,81]. Thus, the relatively high N content of the organs of fertilized plants likely reflects the accumulation of these N compounds.

The two amino acids that dominate the free amino acid pool when plants are supplied with excess N are glutamine and arginine [77,82]. Arginine, which has a very low C:N ratio (6:4), has been implicated as an important product for the accumulation of excess N, possibly as a buffering mechanism against NH<sub>4</sub><sup>+</sup> toxicity [45,83,84]. Again, the accumulation of high levels of arginine in NH<sub>4</sub><sup>+</sup>-fed plants fits with the pattern observed in the G1, and particularly the G2 plants. The very high levels of N and low levels of C in the stalks of the fertilized plants (compared to the control) suggests that the stalk was the most important accumulator for metabolites produced from excess N.

A notable exception to the pattern of increased N with fertilization is the grain, for which there was no significant difference in N content between treatments (Table 4). Our results suggest that at different levels of N supply and plant N content, there was no preferential allocation of accumulated N to the grain, and N that was absorbed post-silking was probably not allocated to the grain. A similar pattern was observed by Ma and Dwyer [85], although it is important to bear in mind the variability among maize hybrids in N metabolism during grain filling [86].

As plants progress through various stages of growth, their uptake, metabolism and partitioning of N may change dramatically. In maize, a significant portion (45–65%) of the grain N is obtained from endogenous N reallocated primarily from the stalk and leaves, while the remaining grain N is obtained from uptake of exogenous soil N [87–90]. Leaf N content at 75 d and 115 d varied as a function of the amount of guano applied (ie. C<G1≤G2), although this was not the case for leaves sampled at 30 d, where there was no clear relationship between quantity of fertilizer applied and leaf N content (Figure 5a). This likely reflects both a reliance on stored seed N early in growth, and the short period of growth prior to transplanting (7 d) during which no fertilizer N was available.

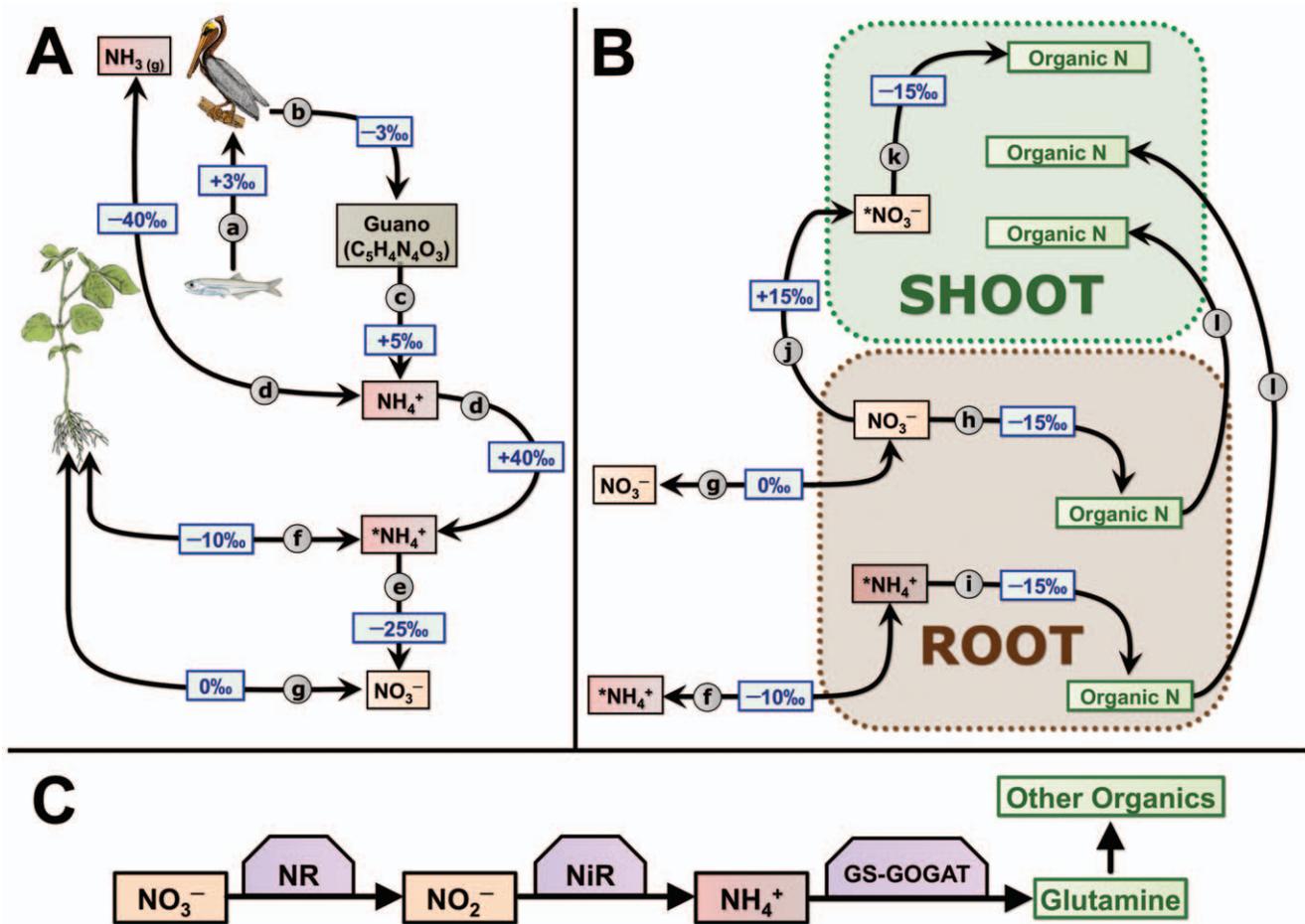
We observed decreases in leaf N content over time, with leaf N content decreasing by 77.9% (C0), 46.9% (G1) and 47.1% (G2) between 30 and 115 d (Figure 5a). The maintenance of very high levels of N in G1 and G2 relative to C0 plants suggests the accumulation of plant N as a result of excess source N [91].

Based on the results of this study, seabird guano fertilization has the potential to significantly alter the C and N economy of maize plants. Specifically, fertilization results in increased N and decreased C:N ratio in plant tissues, which likely arises because of increased accumulation of N-rich metabolites such as arginine, glutamine, and proteins related to NH<sub>4</sub><sup>+</sup> metabolism.

### Intraplant Variation in $\delta^{15}\text{N}$

Intraplant variability in nitrogen isotopic composition for all treatments was large, with maximum differences between mean organ  $\delta^{15}\text{N}$  being 3.0‰ for C0, 12.9‰ for G1 and 11.4‰ for G2 (Figure 4). We found significant differences in the  $\delta^{15}\text{N}$  values between maize plant organs for both control ( $F_{4,20} = 7.41$ ,  $p < 0.001$ ) and fertilized ( $F_{4,20} = 18.60$ ,  $p < 0.001$  for G1;  $F_{4,20} = 28.73$ ,  $p < 0.001$  for G2) treatments (Figure 4). In all treatments, the grain possessed the lowest  $\delta^{15}\text{N}$  value, while anthers had the highest  $\delta^{15}\text{N}$  values in the control treatment and the second-highest  $\delta^{15}\text{N}$  values in the fertilized treatments, following stalks (Figure 4).

Significant variability in  $\delta^{15}\text{N}$  within plants has been recorded in several studies [30,92–98]. Evans [99] suggests that, in general,



**Figure 3. Simplified schematic of fractionation factors associated with decomposition and uptake of seabird guano N.** **A**) Simplified pathway for guano-derived nitrogen. (a) Incorporation of dietary N into consumer tissue N. Tissue–diet fractionation for birds has been calculated to be ~3‰ for most tissues [160,161]. (b) Excretion of dietary N as uric acid. Wainright et al. [162] found bulk guano to be depleted of  $^{15}\text{N}$  by 2.5‰ relative to seabird blood. Moreover, Mizutani et al. [75,163] and Bird et al. [164] found  $\delta^{15}\text{N}$  of uric acid to be very similar to bulk guano  $\delta^{15}\text{N}$ . (c) Conversion of uric acid to  $\text{NH}_4^+$ , according to the experiment performed by Mizutani et al. [163]. (d) Ammonia volatilization. Many studies have found this process to be associated with a large equilibrium fractionation that concentrates  $^{15}\text{N}$  in the remaining substrate ( $^*\text{NH}_4^+$  in the diagram) [54,75,165]. (e) Nitrification. The fractionation factor for the entire process of nitrification in the soil ( $\text{NH}_4^+ \rightarrow \text{NO}_2^- \rightarrow \text{NO}_3^-$ ) is estimated to be between -12 and -35‰ [38,166,167]. (f) Uptake of  $\text{NH}_4^+$  is associated with a nitrogen isotope fractionation ranging from -6 to -30‰ and appears to depend on the concentration of the source  $\text{NH}_4^+$  [33,168]. (g) Uptake of  $\text{NO}_3^-$  by the plant does not appear to be associated with any fractionation [33,169,170]. Both  $\text{NO}_3^-$  and  $\text{NH}_4^+$  may be effluxed from the plant, passively and in some cases actively [171]. **B**) (h)  $\text{NO}_3^-$  assimilation into organic N occurs in the root by the NR-NiR (nitrate reductase-nitrite reductase) and GS-GOGAT (glutamine synthetase–glutamine:oxoglutarate aminotransferase) pathways (see Figure 4C). The reduction of  $\text{NO}_3^-$  to  $\text{NH}_4^+$  is associated with a fractionation factor of -15‰ [37,172]. (i)  $\text{NH}_4^+$  assimilation occurs in the root via the GS-GOGAT pathway and is associated with a fractionation factor of -10 to -15‰ [40,94]. (j, k)  $\text{NO}_3^-$  may also be mobilized to the shoot for assimilation. In this case, this  $\text{NO}_3^-$  pool has already been exposed to  $\text{NO}_3^-$  assimilation in the root and is enriched in  $^{15}\text{N}$  [95]. Therefore, organic N formed from  $\text{NO}_3^-$  in the shoot ( $^*\text{NO}_3^-$ ) will have a higher  $\delta^{15}\text{N}$  value than organic N formed from  $\text{NO}_3^-$  in the root. (l) Organics may be moved between the root and shoot. **C**) Simplified schematic for the assimilation of N by plants. For a more detailed description see Miller and Cramer [171]. All fractionation factors are approximate values representing medians of ranges, which may be large (see text for discussion). doi:10.1371/journal.pone.0033741.g003

plants with  $\text{NO}_3^-$  as the primary N source are characterized by significant intraplant variability, while this is not true for plants with  $\text{NH}_4^+$  as their primary N source. This general pattern results largely from the fact that  $\text{NH}_4^+$  is assimilated into organic N only in the root, while  $\text{NO}_3^-$  assimilation occurs both in roots and shoots (Figure 3B) [57,100,101]. Therefore, organics derived from  $\text{NH}_4^+$  are assimilated from the same N pool in the roots, while  $\text{NO}_3^-$  that has been translocated to the shoot prior to assimilation has already undergone some fractionation (in the roots) and is thus enriched in  $^{15}\text{N}$  [30,95,99].

The  $\delta^{15}\text{N}$  values of the roots were intermediate compared to other above-ground tissues, which does not fit with the scenario described above for  $\text{NO}_3^-$  fed plants in which shoot tissues have

higher  $\delta^{15}\text{N}$  values than roots. In the C0 and G1 plants, the roots did not differ significantly from stalks, grains, or leaves in terms of  $\delta^{15}\text{N}$  (Table 6). In the G2 plants, root  $\delta^{15}\text{N}$  was significantly lower relative to the stalk, but significantly higher than the leaf or grain (Table 6). The lack of a consistent pattern of root vs. shoot  $\delta^{15}\text{N}$  observed in this study likely reflects complex N metabolism, with relative reliance on  $\text{NH}_4^+$  and  $\text{NO}_3^-$ , as well as guano-derived N changing over time.

The relatively low grain  $\delta^{15}\text{N}$  values observed in this study are indicative of the reallocation of stored N. Choi et al. [102] also observed that grain tended to be depleted of  $^{15}\text{N}$  compared to stalks and leaves. This can be attributed to a kinetic isotope effect associated with catabolism and remobilization of stored plant N,

**Table 3.** Isotopic and elemental compositions of plant samples (mean $\pm$ 1 $\sigma$ ).

Treatment	Tissue	Sampling Date	$\delta^{15}\text{N}$ (‰, AIR)	$\delta^{13}\text{C}$ (‰, VPDB)	%N	%C
C0	Leaf	30 d	-0.3 $\pm$ 3.6	-	5.1 $\pm$ 0.7	-
	Leaf	75 d	5.5 $\pm$ 1.4	-	2.7 $\pm$ 0.6	-
	Leaf	115 d	3.9 $\pm$ 1.0	-15.7 $\pm$ 0.8	1.1 $\pm$ 0.3	39.5 $\pm$ 2.3
	Grain	115 d	2.7 $\pm$ 0.7	-12.9 $\pm$ 0.1	2.4 $\pm$ 0.3	42.5 $\pm$ 1.4
	Anther	75 d	5.7 $\pm$ 0.4	-14.0 $\pm$ 0.4	2.4 $\pm$ 0.4	47.0 $\pm$ 2.3
	Root	115 d	3.6 $\pm$ 0.9	-15.0 $\pm$ 0.5	0.8 $\pm$ 0.2	44.7 $\pm$ 1.1
	Stalk	115 d	3.2 $\pm$ 1.4	-15.2 $\pm$ 0.2	1.0 $\pm$ 0.5	47.0 $\pm$ 1.8
	G1	Leaf	30 d	-5.0 $\pm$ 10.0	-	4.9 $\pm$ 0.9
Leaf		75 d	32.4 $\pm$ 2.2	-	4.1 $\pm$ 0.3	-
Leaf		115 d	26.8 $\pm$ 2.0	-15.1 $\pm$ 0.5	2.6 $\pm$ 0.7	40.6 $\pm$ 4.7
Grain		115 d	25.5 $\pm$ 1.6	-14.1 $\pm$ 0.8	2.5 $\pm$ 0.2	45.8 $\pm$ 2.8
Anther		75 d	34.2 $\pm$ 3.4	-13.5 $\pm$ 0.4	3.2 $\pm$ 0.1	48.4 $\pm$ 2.1
Root		115 d	33.1 $\pm$ 4.1	-15.2 $\pm$ 0.3	1.4 $\pm$ 0.7	44.2 $\pm$ 4.2
Stalk		115 d	38.4 $\pm$ 1.9	-15.4 $\pm$ 0.5	2.9 $\pm$ 0.8	36.4 $\pm$ 2.6
G2		Leaf	30 d	6.0 $\pm$ 4.3	-	5.7 $\pm$ 0.4
	Leaf	75 d	38.2 $\pm$ 0.9	-	4.8 $\pm$ 0.2	-
	Leaf	115 d	33.3 $\pm$ 2.7	-15.4 $\pm$ 1.0	3.2 $\pm$ 0.6	42.6 $\pm$ 3.9
	Grain	115 d	33.1 $\pm$ 2.8	-13.3 $\pm$ 0.3	2.6 $\pm$ 0.2	44.6 $\pm$ 1.7
	Anther	75 d	41.8 $\pm$ 2.6	-13.5 $\pm$ 0.4	3.3 $\pm$ 0.5	45.2 $\pm$ 4.4
	Root	115 d	40.1 $\pm$ 2.6	-14.5 $\pm$ 0.6	2.1 $\pm$ 0.7	41.4 $\pm$ 1.8
	Stalk	115 d	44.7 $\pm$ 0.8	-14.7 $\pm$ 0.6	3.4 $\pm$ 0.2	29.7 $\pm$ 0.8

doi:10.1371/journal.pone.0033741.t003

which discriminates against  $^{15}\text{N}$  [103]. The high  $\delta^{15}\text{N}$  values of stalks suggest that this organ is an important source of accumulated N that is remobilized during grain filling. This supports the findings of Ta [104], who found that maize stalks functioned as a significant temporary storage reservoir for N-

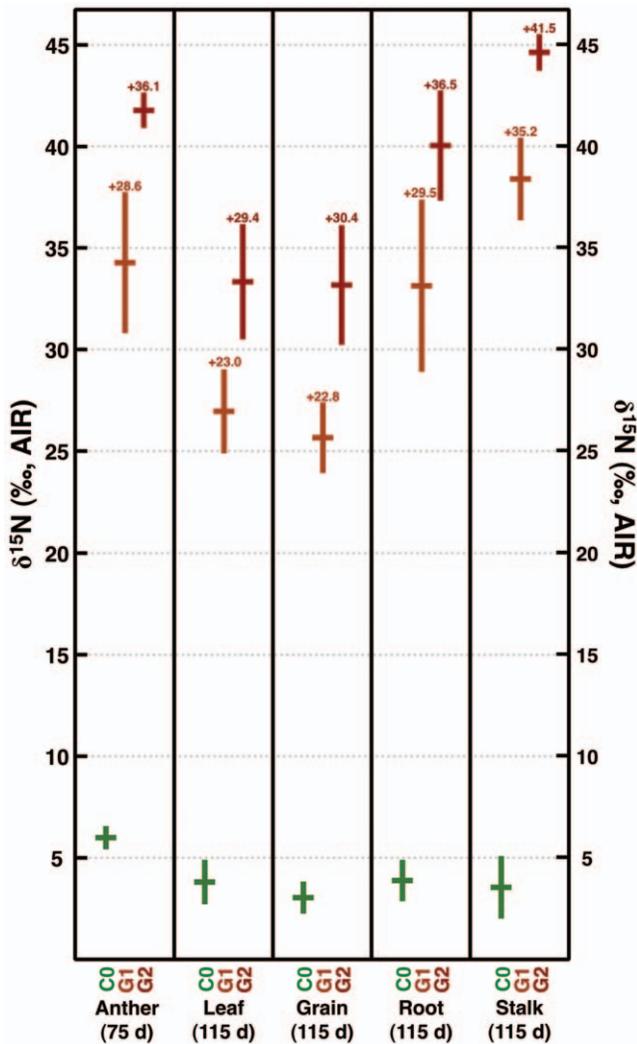
compounds. It is surprising that the leaves at 115 d are not characterized by higher  $\delta^{15}\text{N}$  values in comparison to the grain, as they are thought to be a significant contributor to grain N [105,106]; this is discussed in more detail below. The importance of stalk, compared to leaf, N during grain filling may be specific to

**Table 4.** Results of ANOVA for differences in isotopic and elemental tissue compositions between treatments.

Tissue	Treatment	G1				G2			
		$\delta^{15}\text{N}$ (‰, AIR)	$\delta^{13}\text{C}$ (‰, VPDB)	%N	%C	$\delta^{15}\text{N}$ (‰, AIR)	$\delta^{13}\text{C}$ (‰, VPDB)	%N	%C
Leaf 30 d	C0	0.707	-	0.889	-	0.096	-	0.347	-
	G1	-	-	-	-	0.171	-	0.176	-
Leaf 75 d	C0	<b>&lt;0.001</b>	-	<b>0.008</b>	-	<b>&lt;0.001</b>	-	<b>0.002</b>	-
	G1	-	-	-	-	<b>&lt;0.001</b>	-	<b>0.002</b>	-
Leaf 115 d	C0	<b>&lt;0.001</b>	0.509	<b>0.003</b>	0.884	<b>&lt;0.001</b>	0.857	<b>&lt;0.001</b>	0.414
	G1	-	-	-	-	<b>0.001</b>	0.819	0.357	0.686
Stalk	C0	<b>&lt;0.001</b>	0.640	<b>&lt;0.001</b>	<b>&lt;0.001</b>	<b>&lt;0.001</b>	0.249	<b>&lt;0.001</b>	<b>&lt;0.001</b>
	G1	-	-	-	-	<b>&lt;0.001</b>	0.056	0.415	<b>&lt;0.001</b>
Grain	C0	<b>&lt;0.001</b>	0.066	0.760	0.092	<b>&lt;0.001</b>	0.150	0.463	0.348
	G1	-	-	-	-	<b>&lt;0.001</b>	0.221	0.869	0.632
Anther	C0	<b>&lt;0.001</b>	0.118	<b>0.010</b>	0.746	<b>&lt;0.001</b>	0.135	<b>0.006</b>	0.669
	G1	-	-	-	-	<b>0.017</b>	0.997	0.940	0.280
Root	C0	<b>&lt;0.001</b>	0.746	0.227	0.958	<b>&lt;0.001</b>	0.249	<b>&lt;0.001</b>	0.163
	G1	-	-	-	-	<b>&lt;0.001</b>	0.076	0.415	0.250

Values in boldface are statistically significant ( $p < 0.05$ ).

doi:10.1371/journal.pone.0033741.t004



**Figure 4. Mean nitrogen isotope composition of maize organs; horizontal bars represent means, vertical bars represent standard deviations.** Values above G1 and G2 maize represent differences in nitrogen isotopic composition relative to C0 maize. doi:10.1371/journal.pone.0033741.g004

this variety of maize. Further study of the nitrogen metabolism of different maize hybrids is needed to clarify this issue.

### Temporal Variation in Plant $\delta^{15}\text{N}$ Values

There was significant variability in maize leaves over the course of the experiment (Figure 5B). Maize leaves sampled at 115 d had lower  $\delta^{15}\text{N}$  values than those sampled at 75 d for all treatments; these differences were statistically significant for the fertilized groups, but not for the control group (Table 4). For all treatments, leaf  $\delta^{15}\text{N}$  values were significantly lower at 30 d compared to 75 d (Table 4).

Several studies have attempted to document changes in plant  $\delta^{15}\text{N}$  values over time and/or arising from natural leaf senescence. Kolb and Evans [97] and Garten [107] found no significant differences in the  $\delta^{15}\text{N}$  values of living and abscised leaves, which suggested a lack of  $^{15}\text{N}$  discrimination with N remobilization. Conversely, several other studies have found older or senescent plant leaves to be characterized by higher  $\delta^{15}\text{N}$  values, which has been attributed to a kinetic isotopic fractionation associated with N

catabolism and reallocation [108–110]. We observed no significant difference between leaf  $\delta^{15}\text{N}$  at 75 d and 115 d for the control group, suggesting that under normal circumstances, there is no significant fractionation associated with N remobilization from leaves for this variety of maize. That there was a concurrent decrease in N content and  $\delta^{15}\text{N}$  for leaves between 75 and 115 d in the fertilized plants is counterintuitive, as the reallocation of leaf N to the grain should result in a  $^{15}\text{N}$ -enriched leaf. As was previously suggested for the stalk, we suspect that a significant portion of the leaf N pool consisted of accumulated N in the form of free amino acids (especially arginine and glutamine) as a result of high N supply and, in particular, high source  $[\text{NH}_4^+]$ . The reason that older or senescent plant parts are characterized by higher  $\delta^{15}\text{N}$  values is because the metabolic processes involved (e.g. deamination, transamination) are associated with large kinetic fractionations that concentrate the remaining substrate in  $^{15}\text{N}$  [111]. Therefore, if the majority of the decrease in leaf N between 75 and 115 d is the result of the transfer of organic N products (amino acids) to another part in the plant (e.g. the stalk), which is not associated with any known  $^{15}\text{N}$  fractionation [112], this would help to explain why the leaves are not relatively enriched in  $^{15}\text{N}$  at 115 compared to 75 d.

Leaf  $\delta^{15}\text{N}$  values were more variable at 30 d than at either 75 or 115 d (Figure 5c). This is likely a result of variable reliance on stored and absorbed N sources. Kolb and Evans [97] found that young leaves (*Quercus* and *Encelia*) had an isotopic composition ( $\delta^{15}\text{N}$ ) that reflected both stored and absorbed N, while mature leaf  $\delta^{15}\text{N}$  values reflected primarily absorbed N. Very low leaf  $\delta^{15}\text{N}$  values ( $-12.4$ ,  $-12.4$ ,  $-10.2\%$ ) were observed at 30 d for three of the guano-fertilized maize plants. These compositions probably arise from physiological responses to high soil  $[\text{NH}_4^+]$ . At high extracellular  $[\text{NH}_4^+]$ , influx of  $\text{NH}_4^+$  occurs only via a low-affinity transport system, with high-affinity transport system proteins being down-regulated; this process occurs in concert with the active efflux of  $\text{NH}_4^+$  from the roots [34]. Yoneyama et al. [33] suggest that when  $\text{NH}_4^+$  assimilation is slow (because extracellular  $[\text{NH}_4^+]$  is high),  $\text{NH}_4^+$ -N isotopic fractionation is larger, with relatively more  $^{15}\text{N}$ -enriched  $\text{NH}_4^+$  being effluxed from the cell. Ariz et al. [34] found plants that were most sensitive to  $\text{NH}_4^+$  toxicity also had the lowest tissue  $\delta^{15}\text{N}$  values. The fact that not all plants in the present study were characterized by low leaf  $\delta^{15}\text{N}$  values is difficult to explain, but may be the result of heterogeneous distribution of the guano throughout the soil or genotypic variability in resilience to  $\text{NH}_4^+$  toxicity.

Temporal patterns in plant  $\delta^{15}\text{N}$  values are complicated and are determined by a number of factors. We suspect that significant changes in the N source occurred over time as a result of soil nitrification, and there were also significant changes in [source N] over time. This complication, however, is a reality of working with animal fertilizers, rather than hydroponic solutions, and must be taken into account when interpreting data from field settings.

### Guano Fertilization and Plant Carbon Isotopic Composition

We observed no difference in plant  $\delta^{13}\text{C}$  values resulting from guano fertilization for any of the organs analyzed (Tables 3, 4). In earlier studies, variable plant N sources have been associated with small, but significant variations in  $\delta^{13}\text{C}$  values [113]. It is thought that this association arises because different N sources (and different N source concentrations) may alter plant water-use efficiency and thus change the carbon isotope composition of plant tissues [114].

Previous studies have found plant  $\delta^{13}\text{C}$  values to be distinct in organic vs. inorganic fertilization regimens, an outcome ascribed

**Table 5.** Summary of studies examining the influence of organic fertilization on plant  $\delta^{15}\text{N}$  values.

Fertilizer	Fertilizer $\delta^{15}\text{N}$ (‰, AIR)	Plant	Plant $\delta^{15}\text{N}$ (‰, AIR)	$\Delta^{15}\text{N}_{\text{fertilized-control}}$	Reference
Pig manure	13.9	Maize	7.7	+1.1	[102]
Various composts	17.4±1.2	Maize	17.7	+13.5	[196]
Various composts	17.4±1.2	Nightshade	13.4	+10.7	[196]
Various composts	17.4±1.2	Pepper	14.5	+9.8	[196]
Various composts	17.4±1.2	Mustard	16.3	+12.7	[196]
Various composts	17.4±1.2	Melon	13.3	+10.1	[196]
Various composts	17.4±1.2	Lettuce	13.5	+9.4	[196]
Various composts	17.4±1.2	Spinach	9.5	+3.9	[196]
Various composts	17.4±1.2	Beefsteak plant	19.9	+15.4	[196]
Various composts	17.4±1.2	Sesame	17.8	+12.1	[196]
Pig manure	16.9	Chrysanthemum	10.3	+3.5	[194]
Pig manure	16.9	Cabbage	13.3	+5.6	[194]
Sheep manure	–	Sweet pepper	10.0	–	[198]
Chicken manure	–	Sweet pepper	10.2	–	[198]
Horse manure	–	Sweet pepper	9.8	–	[198]
Livestock manure	8.7±0.2	Orange (pulp)	9.0	–	[132]
Poultry manure	8.6±0.3	Orange (pulp)	8.5	–	[132]
Livestock manure	8.7±0.2	Orange (juice)	8.5	–	[132]
Poultry manure	8.6±0.3	Orange (juice)	7.9	–	[132]
Pig manure	16.4	Chinese cabbage	12.5	+11.0	[195]
Mixed (Cattle+poultry manure)	16.7	Tomato	13.5	+10.2	[199]
Mixed (Cattle+poultry manure)	9.9	Tomato	7.9	+4.6	[199]

doi:10.1371/journal.pone.0033741.t005

to higher rates of soil microbiological activity [115,116]. Specifically, Georgi et al. [116] suggest that  $\text{CO}_2$  released during decomposition is depleted of  $^{13}\text{C}$ . Because control and fertilized plants were grown in the same growth chamber, there would be no differences in the  $\delta^{13}\text{C}$  of  $\text{CO}_2$  utilized by either group of plants, although this may not be true for an agricultural field fertilized with guano. In general, the influence on nitrogenous fertilizers (both organic and inorganic) on plant  $\delta^{13}\text{C}$  is unclear. Experimental results have been conflicting, with studies finding  $\delta^{13}\text{C}$  values to increase [117–122], decrease [123], or be unaffected [120,124] in response to N fertilization. The relationship between N fertilizer application and plant  $\delta^{13}\text{C}$  is likely mediated by several factors and warrants further study. We likely did not detect any difference in plant  $\delta^{13}\text{C}$  values resulting from fertilization because the magnitude of difference would be quite small [113] and our sample size was also quite small ( $n = 5$  per treatment).

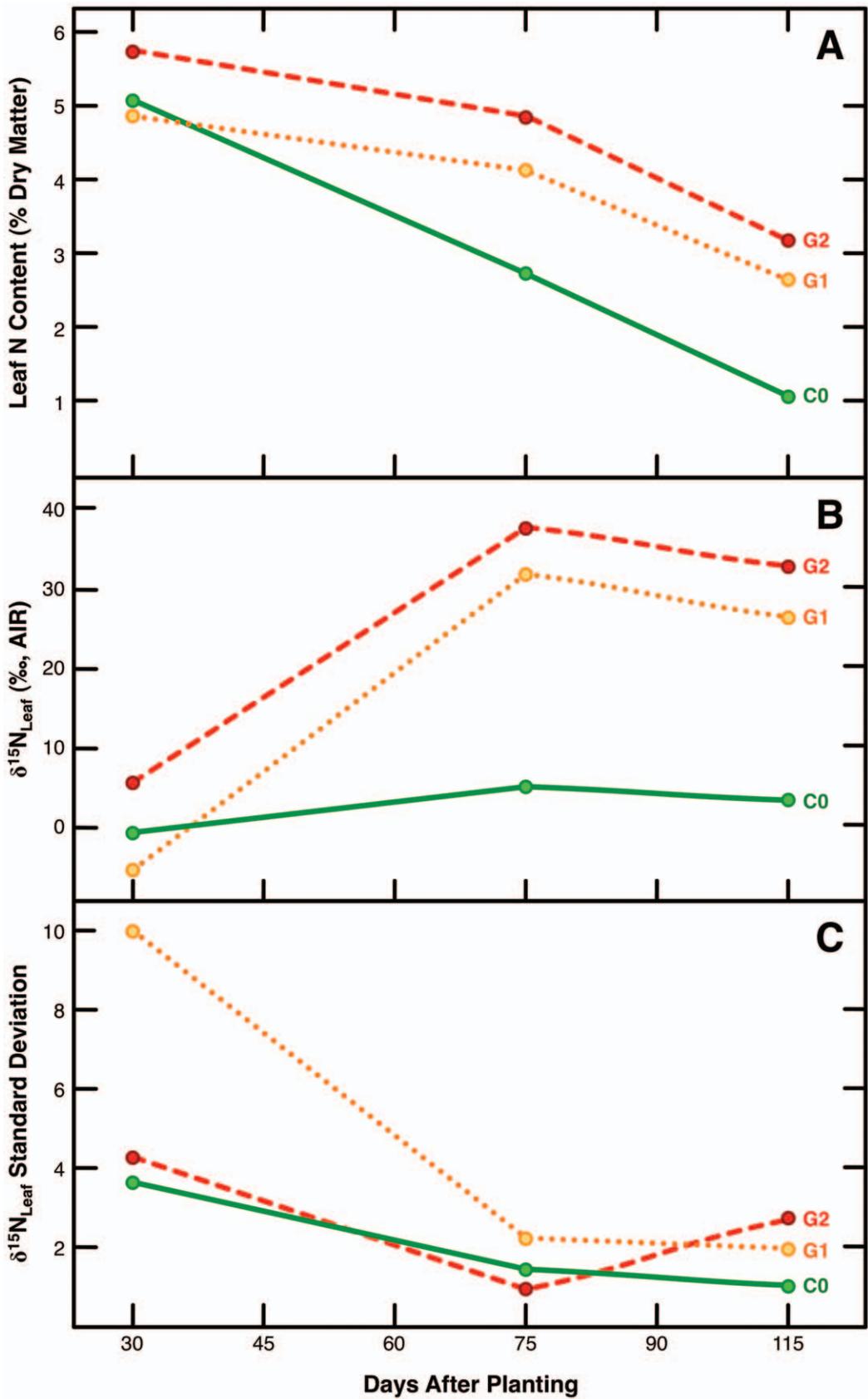
### Implications for Food Chemistry

Seabird guano is becoming increasingly popular as an organic alternative among farmers in the United States and Europe [5]. Moreover, as the demand for organically grown produce soars worldwide [125], there is an increased incentive for farmers in areas in close proximity to guano deposits (e.g. Peru, Ecuador, Chile, and Namibia) to use this fertilizer and market their produce as organic [5]. In recent years, there has been a surge in isotopic research directed at demonstrating isotopic distinctions between conventional and organically grown produce [126–136]. The reason that this technique may sometimes be effective is primarily that inorganic fertilizers tend to have  $\delta^{15}\text{N}$  values close to 0‰, while organic fertilizers tend to have higher  $\delta^{15}\text{N}$  values, although

there is great variability (Table 2). Based on the results of this study, the application of seabird guano in an organic fertilization regime would result in a very large  $^{15}\text{N}$  enrichment of all plant tissues in comparison to unfertilized plants, or to plants treated with chemical fertilizers. The magnitude of this difference is much greater than what has been observed for other organic fertilizers (Table 5), and thus isotopic data would be useful in verifying use of seabird guano. Moreover, the very high  $\delta^{15}\text{N}$  value of the guano itself suggests that its presence in mixed organic fertilizers should also be detectable via isotope ratio mass spectrometry.

### Implications for Archaeology

Stable isotope analysis ( $\delta^{13}\text{C}$  and  $\delta^{15}\text{N}$  in particular) plays an increasingly important role in the reconstruction of prehistoric diet. Dietary reconstruction requires a thorough understanding of the sources of isotopic variation in the foods that were consumed [137]. Recently, the notion that animal manure may have influenced the  $\delta^{15}\text{N}$  values of plants grown in prehistoric Europe has been proposed [138–140] and integrated into regional paleodietary studies. In the Andean region, several fertilizers are thought to have been of some importance in prehispanic agriculture including llama dung [141] and seabird guano [9,142,143]. Based on the large settlements that developed on the coast of Peru (e.g. Moche, Chimú) and the relative infertility of local soils, Nordt et al. [8] have suggested that the application of some kind of nitrogenous fertilizer, possibly seabird guano, would have been necessary to maintain agricultural productivity in at least some parts of the region. Direct evidence for fertilization, however, is very difficult to come by. One of the primary goals of this study was to determine whether or not the enrichment in  $^{15}\text{N}$



**Figure 5. Temporal patterns in isotopic and element composition. (A)** Leaf N content, **(B)** Leaf  $\delta^{15}\text{N}$ , and **(C)** standard deviation for Leaf  $\delta^{15}\text{N}$ . doi:10.1371/journal.pone.0033741.g005

**Table 6.** Results of ANOVA for differences in nitrogen isotopic composition between plant parts.

Treatment	Tissue	Leaf	Anther	Root	Stalk
C0	Grain	0.319	<0.001	0.626	0.908
	Leaf	–	0.041	0.981	0.803
	Anther	–	–	0.013	0.004
	Root	–	–	–	0.980
G1	Grain	0.915	0.017	0.077	<0.001
	Leaf	–	0.035	0.152	<0.001
	Anther	–	–	0.999	0.309
	Root	–	–	–	0.252
G2	Grain	0.999	<0.001	<0.001	<0.001
	Leaf	–	<0.001	<0.001	<0.001
	Anther	–	–	0.709	0.259
	Root	–	–	–	0.022

Values in boldface are statistically significant ( $p < 0.05$ ).  
doi:10.1371/journal.pone.0033741.t006

resulting from guano fertilization would be sufficient to detect this agricultural practice in the isotopic composition of a human or animal consuming the fertilized plant. Based on the results of this study and others that have examined the biogeochemistry of seabird-associated sites (summarized in Table 2), the application of seabird guano to agricultural fields would have caused a significant increase in the  $\delta^{15}\text{N}$  value of plants and of animals consuming these plants. In archaeological bone collagen from western South America, high  $\delta^{15}\text{N}$  values are usually accompanied by high  $\delta^{13}\text{C}$  values. This pattern applies to both humans [144–146] and domestic animals [147], and has generally been attributed to the

consumption of high trophic-level marine resources (e.g. predatory fish, marine mammals). Conversely, this pattern may also be caused by the consumption of maize (a  $\text{C}_4$  plant) fertilized with seabird guano, which appears (isotopically) very much like a high-trophic level marine organism. As such, it is important to be mindful of the possibility of guano-fertilization when interpreting diet, not just on the coast, but in the interior highland region as well. According to ethnohistoric documents, guano was moved great distances and prized by groups living in the highlands as an essential component in maize agriculture [9].

The Andes were certainly not the only region in which seabird guano was used extensively as a fertilizer. Millions of tonnes of guano were exported to Europe and North America during the nineteenth century and Peruvian seabird guano was the most highly prized fertilizer at that time [148–150]. Isotopic analysis is being employed with increased frequency within the context of historical archaeology [151–159], a period during which the possible influence of seabird guano must also be considered.

## Supporting Information

### Table S1 Raw isotopic and elemental data for all samples analyzed.

(XLS)

## Acknowledgments

Steve Bartlett (Biotron), Kim Law and Li Huang (LSIS) provided technical assistance. Sharon Buck assisted with sample preparation. This is Laboratory for Stable Isotope Science Contribution #279.

## Author Contributions

Conceived and designed the experiments: PS FJL JFM CDW. Performed the experiments: PS. Analyzed the data: PS EJJ CDW. Wrote the paper: PS.

## References

- Duffy DC (1994) The guano islands of Peru: the once and future management of a renewable resource. In: Nettleship DN, Burger J, Gochfeld M, eds. Seabirds on Islands, Threats, Case Studies and Action Plans. Cambridge: BirdLife Conservation Series, No. 1. pp 68–76.
- Murphy RC (1981) The guano and the anchoveta fishery. In: Glantz MH, Thompson JD, eds. Resource Management and Environmental Uncertainty: Lessons from Coastal Upwelling Fisheries. New York: Wiley. pp 81–106.
- Mathew WM (1970) Peru and the British Guano Market, 1840–1870. The Economic History Review 23: 112–128.
- Zapata F, Arrillaga JL (2002) Agronomic evaluation of guano sources by means of isotope techniques. In: Sikora F, ed. Assessment of soil phosphorus status and management of phosphatic fertilisers to optimise crop production. Vienna: International Atomic Energy Agency. pp 83–89.
- Romero S (2008) Peru guards its guano as demand soars again. New York Times. May 30 ed. New York.
- Garcilaso de la Vega I (1966) Royal Commentaries of the Incas and General History of Peru. Livermore HL, translator. Austin: University of Texas Press. 1530 p.
- Cieza de León Pd (1964) The Travels of Pedro de Cieza de León: A.D. 1532–50, contained in the first part of his Chronicle of Peru. Markham CR, translator. New York: Franklin. 438 p.
- Nordt L, Hayashida F, Hallmark T, Crawford C (2004) Late prehistoric soil fertility, irrigation management, and agricultural production in northwest coastal Peru. Geoarchaeology 19: 21–46.
- Julien CJ (1985) Guano and resource control in sixteenth-century Arequipa. In: Masuda S, Shimada I, Morris C, eds. Andean Ecology and Civilization: An Interdisciplinary Perspective on Andean Ecological Complementarity. Tokyo: University of Tokyo Press. pp 185–231.
- Powell GVN, Fourqurean JW, Kenworthy WJ, Zieman JC (1991) Bird colonies cause seagrass enrichment in a subtropical estuary: Observational and experimental evidence. Estuarine, Coastal and Shelf Science 32: 567–579.
- Ishizuka K (1966) Ecology of the ornithophilous plant communities on breeding places of the black-tailed gull, *Larus crassirostris*, along the coast of Japan: I. Vegetation analysis. Ecological Review 16: 229–244.
- Hutchinson GE (1950) Survey of Existing Knowledge of Biogeochemistry: 3. The Biogeochemistry of Vertebrate Excretion. Bulletin of the American Museum of Natural History 96: 1–554.
- Anderson WB, Polis GA (1999) Nutrient fluxes from water to land: seabirds affect plant nutrient status on Gulf of California islands. Oecologia 118: 324–332.
- Ryan PG, Watkins BP (1989) The influence of physical factors and ornithogenic products on plant and arthropod abundance at an Inland Nunatak group in Antarctica. Polar Biology 10: 151–160.
- Burger AE, Lindeboom HJ, Williams AJ (1978) The mineral and energy contributions of guano of selected species of birds to the Marion Island terrestrial ecosystem. South African Journal of Antarctic Research 8: 59–70.
- Engelskjøn T (1986) Botany of two Antarctic mountain ranges: Gjelsvikfjella and Mühlig-Hofmannfjella, Dronning Maud Land. Polar Research 4: 205–224.
- Leentvaar P (1967) Observations in guanotrophic environments. Hydrobiologia 29: 441–489.
- McColl JG, Burger J (1976) Chemical Inputs by a Colony of Franklin's Gulls Nesting in Cattails. American Midland Naturalist 96: 270–280.
- Smith VR (1978) Animal-plant-soil nutrient relationships on Marion Island (Subantarctic). Oecologia 32: 239–253.
- Spir TW, Cowling JC (1984) Ornithogenic soils of the Cape Bird Adeline penguin rookeries, Antarctica: 1. Chemical Properties. Polar Biology 2: 199–205.
- Ellis JC (2005) Marine Birds on Land: A Review of Plant Biomass, Species Richness, and Community Composition in Seabird Colonies. Plant Ecology 181: 227–241.
- Vidal E, Jouventin P, Frenot Y (2003) Contribution of alien and indigenous species to plant-community assemblages near penguin rookeries at Crozet archipelago. Polar Biology 26: 432–437.
- Lindeboom HJ (1984) The Nitrogen Pathway in a Penguin Rookery. Ecology 65: 269–277.
- Williams AJ, Berruti A (1978) Mineral and energy contributions of feathers moulted by penguins, gulls and cormorants to the Marion Island terrestrial ecosystem. South African Journal of Antarctic Research 8: 71–74.

25. Williams AJ, Burger AE, Berruti A (1978) Mineral and energy contributions of carcasses of selected species of seabirds to the Marion Island terrestrial ecosystem. *South African Journal of Antarctic Research* 8: 53–58.
26. Siegfried WR, Williams AJ, Burger AE, Berruti A (1978) Mineral and energy contributions of eggs of selected species of seabirds to the Marion Island terrestrial ecosystem. *South African Journal of Antarctic Research* 8: 75–87.
27. Gillham ME (1956) Ecology of the Pembroke Islands: V. Manuring by the Colonial Seabirds and Mammals, with a Note on Seed Distribution by Gulls. *Journal of Ecology* 44: 429–454.
28. Crawford NM, Glass ADM (1998) Molecular and physiological aspects of nitrate uptake in plants. *Trends in Plant Science* 3: 389–395.
29. Andrews M (1986) The partitioning of nitrate assimilation between root and shoot of higher plants. *Plant, Cell & Environment* 9: 511–519.
30. Yoneyama T, Kaneko A (1989) Variations in the Natural Abundance of  $^{15}\text{N}$  in Nitrogenous Fractions of Komatsuna Plants Supplied with Nitrate. *Plant and Cell Physiology* 30: 957–962.
31. Mariotti A, Mariotti F, Amargor N, Pizelle G, Ngambi JM, et al. (1980) Fractionnements isotopiques de l'azote lors des processus d'absorption des nitrates et de fixation de l'azote atmosphérique par les plantes. *Physiologie Végétale* 18: 163–181.
32. Yoneyama T, Ito O, Engelaar WMHG (2003) Uptake, metabolism and distribution of nitrogen in crop plants traced by enriched and natural  $^{15}\text{N}$ : Progress over the last 30 years. *Phytochemistry Reviews* 2: 121–132.
33. Yoneyama T, Matsumaru T, Usui K, Engelaar WMHG (2001) Discrimination of nitrogen isotopes during absorption of ammonium and nitrate at different nitrogen concentrations by rice (*Oryza sativa* L.) plants. *Plant, Cell & Environment* 24: 133–139.
34. Ariz I, Cruz C, Moran J, González-Moro M, García-Olaverri C, et al. (2011) Depletion of the heaviest stable N isotope is associated with  $\text{NH}_4^+/\text{NH}_3$  toxicity in  $\text{NH}_4^+$ -fed plants. *BMC Plant Biology* 11: 83.
35. Flores-Delgado L, Fedick SL, Solleiro-Rebolledo E, Palacios-Mayorga S, Ortega-Larrocea P, et al. (2011) A sustainable system of a traditional precision agriculture in a Maya homestead: Soil quality aspects. *Soil and Tillage Research* 113: 112–120.
36. Pritchard ES, Guy RD (2005) Nitrogen isotope discrimination in white spruce fed with low concentrations of ammonium and nitrate. *Trees - Structure and Function* 19: 89–98.
37. Ledgard SF, Woo KC, Bergersen FJ (1985) Isotopic fractionation during reduction of nitrate and nitrite by extracts of spinach leaves. *Australian Journal of Plant Physiology* 12: 631–640.
38. Robinson D (2001)  $\delta^{15}\text{N}$  as an integrator of the nitrogen cycle. *Trends in Ecology & Evolution* 16: 153–162.
39. Glass ADM, Brito DT, Kaiser BN, Kronzucker HJ, Kumar A, et al. (2001) Nitrogen transport in plants, with an emphasis on the regulation of fluxes to match plant demand. *Journal of Plant Nutrition and Soil Science* 164: 199–207.
40. Yoneyama T, Kamachi K, Yamaya T, Mae T (1993) Fractionation of Nitrogen Isotopes by Glutamine Synthetase Isolated from Spinach Leaves. *Plant and Cell Physiology* 34: 489–491.
41. Subedi KD, Ma BL (2005) Nitrogen Uptake and Partitioning in Stay-Green and Leafy Maize Hybrids. *Crop Science* 45: 740–747.
42. Ishida A (1997) Seed germination and seedling survival in a colony of the common cormorant, *Phalacrocorax carbo*. *Ecological Research* 12: 249–256.
43. Mulder C, Keall S (2001) Burrowing seabirds and reptiles: impacts on seeds, seedlings and soils in an island forest in New Zealand. *Oecologia* 127: 350–360.
44. Bremner JM, Krogmeier MJ (1989) Evidence that the adverse effect of urea fertilizer on seed germination in soil is due to ammonia formed through hydrolysis of urea by soil urease. *Proceedings of the National Academy of Sciences* 86: 8185–8188.
45. Roosta HR, Schjoerring JK (2007) Effects of Ammonium Toxicity on Nitrogen Metabolism and Elemental Profile of Cucumber Plants. *Journal of Plant Nutrition* 30: 1933–1951.
46. Schortemeyer M, Stamp P, Feil B (1997) Ammonium Tolerance and Carbohydrate Status in Maize Cultivars. *Annals of Botany* 79: 25–30.
47. Wootton JT (1991) Direct and indirect effects of nutrients on intertidal community structure: variable consequences of seabird guano. *Journal of Experimental Marine Biology and Ecology* 151: 139–153.
48. Weseloh DV, Brown RT (1971) Plant Distribution within a Heron Rookery. *American Midland Naturalist* 86: 57–64.
49. Clarkson DT, Scattergood CB (1982) Growth and Phosphate Transport in Barley and Tomato Plants During the Development of, and Recovery from, Phosphate-stress. *Journal of Experimental Botany* 33: 865–875.
50. Ligeza S, Smal H (2003) Accumulation of nutrients in soils affected by perennial colonies of piscivorous birds with reference to biogeochemical cycles of elements. *Chemosphere* 52: 595–602.
51. García LV, Marañón T, Ojeda F, Clemente L, Redondo R (2002) Seagull influence on soil properties, chenopod shrub distribution, and leaf nutrient status in semi-arid Mediterranean islands. *Oikos* 98: 75–86.
52. Young HS, McCauley DJ, Dirzo R (2011) Differential responses to guano fertilization among tropical tree species with varying functional traits. *American Journal of Botany* 98: 207–214.
53. Mizota C (2009) Temporal variations in the concentration and isotopic signature of ammonium- and nitrate-nitrogen in soils under a breeding colony of Black-tailed Gulls (*Larus crassirostris*) on Kabushima Island, northeastern Japan. *Applied Geochemistry* 24: 328–332.
54. Mizutani H, Hasegawa H, Wada E (1986) High nitrogen isotope ratio for soils of seabird rookeries. *Biogeochemistry* 2: 221–247–247.
55. Schmidt S, Dennison WC, Moss GJ, Stewart GR (2004) Nitrogen ecophysiology of Heron Island, a subtropical coral cay of the Great Barrier Reef, Australia. *Functional Plant Biology* 31: 517–528.
56. Wait DA, Aubrey DP, Anderson WB (2005) Seabird guano influences on desert islands: soil chemistry and herbaceous species richness and productivity. *Journal of Arid Environments* 60: 681–695.
57. Raven JA, Smith FA (1976) Nitrogen assimilation and transport in vascular land plants in relation to intracellular pH regulation. *New Phytologist* 76: 415–431.
58. Deignan MT, Lewis OAM (1988) The inhibition of ammonium uptake by nitrate in wheat. *New Phytologist* 110: 1–3.
59. Guo S, Kaldenhoff R, Uehlein N, Sattelmacher B, Brueck H (2007) Relationship between water and nitrogen uptake in nitrate- and ammonium-supplied *Phaseolus vulgaris* L. plants. *Journal of Plant Nutrition and Soil Science* 170: 73–80.
60. Britto DT, Siddiqi MY, Glass ADM, Kronzucker HJ (2001) Futile transmembrane  $\text{NH}_4^+$  cycling: A cellular hypothesis to explain ammonium toxicity in plants. *Proceedings of the National Academy of Sciences* 98: 4255–4258.
61. Gillham ME (1960) Destruction of indigenous heath vegetation in Victorian sea-bird colonies. *Australian Journal of Botany* 8: 277–317.
62. Hartz TK, Johnstone PR (2006) Nitrogen availability from high-nitrogen-containing organic fertilizers. *HortTechnology* 16: 39–42.
63. Mizutani H, Kabaya Y, Wada E (1991) Linear correlation between latitude and soil  $^{15}\text{N}$  enrichment at seabird rookeries. *Naturwissenschaften* 78: 34–36.
64. McNabb EMA, McNabb RA, Prather ID, Conner RN, Adkisson CS (1980) Nitrogen Excretion by Turkey Vultures. *The Condor* 82: 219–223.
65. Mizutani H, Wada E (1985) High-performance liquid chromatographic determination of uric acid in soil. *Journal of Chromatography A* 331: 359–369.
66. Gaskell M, Smith R (2007) Nitrogen sources for organic vegetable crops. *HortTechnology* 17: 431–441.
67. Staunton Smith J, Johnson CR (1995) Nutrient inputs from seabirds and humans on a populated coral cay. *Marine Ecology Progress Series* 124: 189–200.
68. Hadas A, Rosenberg R (1992) Guano as a nitrogen source for fertigation in organic farming. *Nutrient Cycling in Agroecosystems* 31: 209–214.
69. Forero MG, Bortolotti GR, Hobson KA, Donazar JA, Bertelotti M, et al. (2004) High trophic overlap within the seabird community of Argentinean Patagonia: a multiscale approach. *Journal of Animal Ecology* 73: 789–801.
70. Schoeninger MJ, DeNiro MJ (1984) Nitrogen and carbon isotopic composition of bone collagen from marine and terrestrial animals. *Geochimica et Cosmochimica Acta* 48: 625–639.
71. Szpak P, Orchard TJ, Gröcke DR (2009) A Late Holocene vertebrate food web from southern Haida Gwaii (Queen Charlotte Islands, British Columbia). *Journal of Archaeological Science* 36: 2734–2741.
72. Szpak P, Orchard TJ, McKechnie I, Gröcke DR (2012) Historical ecology of late Holocene sea otters (*Enhydra lutris*) from northern British Columbia: isotopic and zooarchaeological perspectives. *Journal of Archaeological Science* 39: 1553–1571.
73. Loder TC, Ganning B, Love JA (1996) Ammonia nitrogen dynamics in coastal rockpools affected by gull guano. *Journal of Experimental Marine Biology and Ecology* 196: 113–129.
74. Kirchmann H (1991) Carbon and nitrogen mineralization of fresh, aerobic and anaerobic animal manures during incubation with soil. *Swedish Journal of Agricultural Research* 21: 165–173.
75. Mizutani H, Kabaya Y, Wada E (1985) Ammonia volatilization and high  $^{15}\text{N}/^{14}\text{N}$  ratio in a penguin rookery in Antarctica. *Geochemical Journal* 19: 323–327.
76. Burger M, Venterea RT (2008) Nitrogen Immobilization and Mineralization Kinetics of Cattle, Hog, and Turkey Manure Applied to Soil. *Soil Science Society of America Journal* 72: 1570–1579.
77. Okano K, Chutani K, Matsuo K (1997) Suitable level of nitrogen fertilizer for tea (*Camellia sinensis* L.) plants in relation to growth, photosynthesis, nitrogen uptake and accumulation of free amino acids. *Japanese Journal of Crop Science* 66: 279–287.
78. Nasholm T, Ericsson A (1990) Seasonal changes in amino acids, protein and total nitrogen in needles of fertilized Scots pine trees. *Tree Physiology* 6: 267–281.
79. Warren CR, Adams MA (2000) Capillary electrophoresis for the determination of major amino acids and sugars in foliage: application to the nitrogen nutrition of sclerophyllous species. *Journal of Experimental Botany* 51: 1147–1157.
80. Ruan J, Gerendás J, Härdter R, Sattelmacher B (2007) Effect of root zone pH and form and concentration of nitrogen on accumulation of quality-related components in green tea. *Journal of the Science of Food and Agriculture* 87: 1505–1516.
81. Garnica M, Houdusse F, Zamarréño AM, García-Mina JM (2010) Nitrate modifies the assimilation pattern of ammonium and urea in wheat seedlings. *Journal of the Science of Food and Agriculture* 90: 357–369.

82. Ruan J, Haerdter R, Gerendás J (2010) Impact of nitrogen supply on carbon/nitrogen allocation: a case study on amino acids and catechins in green tea [*Camellia sinensis* (L.) O. Kuntze] plants. *Plant Biology* 12: 724–734.
83. Smolders AJP, den Hartog C, van Gestel CBL, Roelofs JGM (1996) The effects of ammonium on growth, accumulation of free amino acids and nutritional status of young phosphorus deficient *Stratiotes aloides* plants. *Aquatic Botany* 53: 85–96.
84. Potel F, Valadier M-H, Ferrario-Méry S, Grandjean O, Morin H, et al. (2009) Assimilation of excess ammonium into amino acids and nitrogen translocation in *Arabidopsis thaliana*—roles of glutamate synthases and carbamoylphosphate synthetase in leaves. *FEBS Journal* 276: 4061–4076.
85. Ma BL, Dwyer LM (1998) Nitrogen uptake and use of two contrasting maize hybrids differing in leaf senescence. *Plant and Soil* 199: 283–291.
86. Rajcan I, Tollenaar M (1999) Source: sink ratio and leaf senescence in maize: II. Nitrogen metabolism during grain filling. *Field Crops Research* 60: 255–265.
87. Gallais A, Coque M, Quillére I, Prioul J-L, Hirel B (2006) Modelling post-silking nitrogen fluxes in maize (*Zea mays*) using  $^{15}\text{N}$ -labelling field experiments. *New Phytologist* 172: 696–707.
88. Gallais A, Coque M (2005) Genetic variation and selection for nitrogen use efficiency in maize: A synthesis. *Maydica* 50: 531–547.
89. Tsai CY, Huber DM, Warren HL (1980) A Proposed Role of Zein and Glutelin as N Sinks in Maize. *Plant Physiology* 66: 330–333.
90. Below FE, Christensen LE, Reed AJ, Hageman RH (1981) Availability of Reduced N and Carbohydrates for Ear Development of Maize. *Plant Physiology* 68: 1186–1190.
91. Binford GD, Blackmer AM, El-Hout NM (1990) Tissue test for excess nitrogen during corn production. *Agronomy Journal* 82: 124–129.
92. Dijkstra P, Williamson C, Menyailo O, Doucet R, Koch G, et al. (2003) Nitrogen stable isotope composition of leaves and roots of plants growing in a forest and a meadow. *Isotopes in Environmental and Health Studies* 39: 29–39.
93. Hobbie EA, Macko SA, Williams M (2000) Correlations between foliar  $\delta^{15}\text{N}$  and nitrogen concentrations may indicate plant-mycorrhizal interactions. *Oecologia* 122: 273–283.
94. Yoneyama T, Omata T, Nakata S, Yazaki J (1991) Fractionation of Nitrogen Isotopes during the Uptake and Assimilation of Ammonia by Plants. *Plant and Cell Physiology* 32: 1211–1217.
95. Evans RD, Bloom AJ, Sukrapanna SS, Ehleringer JR (1996) Nitrogen isotope composition of tomato (*Lycopersicon esculentum* Mill. cv. T-5) grown under ammonium or nitrate nutrition. *Plant, Cell & Environment* 19: 1317–1323.
96. Högborg P, Högborg MN, Quist ME, Ekblad ALF, Näsholm T (1999) Nitrogen isotope fractionation during nitrogen uptake by ectomycorrhizal and non-mycorrhizal *Pinus sylvestris*. *New Phytologist* 142: 569–576.
97. Kolb KJ, Evans RD (2002) Implications of leaf nitrogen recycling on the nitrogen isotope composition of deciduous plant tissues. *New Phytologist* 156: 57–64.
98. Högborg P, Högborg L, Schinkel H, Högborg M, Johannisson C, et al. (1996)  $^{15}\text{N}$  abundance of surface soils, roots and mycorrhizas in profiles of European forest soils. *Oecologia* 108: 207–214.
99. Evans RD (2001) Physiological mechanisms influencing plant nitrogen isotope composition. *Trends in Plant Science* 6: 121–126.
100. Lewis OAM, Leidi EO, Lips SH (1989) Effect of nitrogen source on growth response to salinity stress in maize and wheat. *New Phytologist* 111: 155–160.
101. Murphy AT, Lewis OAM (1987) Effect of nitrogen feeding source on the supply of nitrogen from root to shoot and the site of nitrogen assimilation in maize (*Zea mays* L. CV. R201). *New Phytologist* 107: 327–333.
102. Choi W-J, Lee S-M, Ro H-M, Kim K-C, Yoo S-H (2002) Natural  $^{15}\text{N}$  abundances of maize and soil amended with urea and composted pig manure. *Plant and Soil* 245: 223–232.
103. Tcherkez G (2011) Natural  $^{15}\text{N}/^{14}\text{N}$  isotope composition in  $\text{C}_3$  leaves: are enzymatic isotope effects informative for predicting the  $^{15}\text{N}$ -abundance in key metabolites? *Functional Plant Biology* 38: 1–12.
104. Ta CT (1991) Nitrogen Metabolism in the Stalk Tissue of Maize. *Plant Physiology* 97: 1375–1380.
105. Beauchamp EG, Kamenberg LW, Hunter RB (1976) Nitrogen Accumulation and Translocation in Corn Genotypes Following Silking. *Agronomy Journal* 68: 418–422.
106. Donnison IS, Gay AP, Thomas H, Edwards KJ, Edwards D, et al. (2007) Modification of nitrogen remobilization, grain fill and leaf senescence in maize (*Zea mays*) by transposon insertional mutagenesis in a protease gene. *New Phytologist* 173: 481–494.
107. Garten CT (1993) Variation in Foliar  $^{15}\text{N}$  Abundance and the Availability of Soil Nitrogen on Walker Branch Watershed. *Ecology* 74: 2098–2113.
108. Gebauer G, Giesemann A, Schulze E, Jäger H (1994) Isotope ratios and concentrations of sulfur and nitrogen in needles and soils of *Picea abies* stands as influenced by atmospheric deposition of sulfur and nitrogen compounds. *Plant and Soil* 164: 267–281.
109. Näsholm T (1994) Removal of nitrogen during needle senescence in Scots pine (*Pinus sylvestris* L.). *Oecologia* 99: 290–296.
110. Choi W-J, Chang SX, Ro H-M (2005) Seasonal Changes of Shoot Nitrogen Concentrations and  $^{15}\text{N}/^{14}\text{N}$  Ratios in Common Reed in a Constructed Wetland. *Communications in Soil Science and Plant Analysis* 36: 2719–2731.
111. Högborg P (1997) Tansley Review No. 95  $^{15}\text{N}$  natural abundance in soil-plant systems. *New Phytologist* 137: 179–203.
112. Robinson D, Handley LL, Scrimgeour CM (1998) A theory for  $^{15}\text{N}/^{14}\text{N}$  fractionation in nitrate-grown vascular plants. *Planta* 205: 397–406.
113. Yin Z-H, Raven JA (1998) Influences of different nitrogen sources on nitrogen- and water-use efficiency, and carbon isotope discrimination in  $\text{C}_3$  *Triticum aestivum* L. and  $\text{C}_4$  *Zea mays* L. plants. *Planta* 205: 574–580.
114. Raven JA, Wollenweber B, Handley LL (1992) A comparison of ammonium and nitrate as nitrogen sources for photolithotrophs. *New Phytologist* 121: 19–32.
115. Camin F, Perini M, Bontempo L, Fabroni S, Faedi W, et al. (2011) Potential isotopic and chemical markers for characterising organic fruits. *Food Chemistry* 125: 1072–1082.
116. Georgi M, Voerkelius S, Rossmann A, Graßmann J, Schnitzler W (2005) Multielement Isotope Ratios of Vegetables from Integrated and Organic Production. *Plant and Soil* 275: 93–100.
117. Serret M, Ortiz-Monasterio I, Pardo A, Araus J (2008) The effects of urea fertilisation and genotype on yield, nitrogen use efficiency,  $\delta^{15}\text{N}$  and  $\delta^{13}\text{C}$  in wheat. *Annals of Applied Biology* 153: 243–257.
118. Kondo M, Pablico P, Aragones D, Agbisit R (2004) Genotypic variations in carbon isotope discrimination, transpiration efficiency, and biomass production in rice as affected by soil water conditions and N. *Plant and Soil* 267: 165–177.
119. Cabrera-Bosquet L, Molero G, Bort J, Nogués S, Araus JL (2007) The combined effect of constant water deficit and nitrogen supply on WUE, NUE and  $\Delta^{13}\text{C}$  in durum wheat potted plants. *Annals of Applied Biology* 151: 277–289.
120. Jenkinson D, Coleman K, Harkness D (1995) The influence of fertilizer nitrogen and season on the carbon-13 abundance of wheat straw. *Plant and Soil* 171: 365–367.
121. Zhao LJ, Xiao HL, Liu XH (2007) Relationships Between Carbon Isotope Discrimination and Yield of Spring Wheat Under Different Water and Nitrogen Levels. *Journal of Plant Nutrition* 30: 947–963.
122. Iqbal MM, Akhter J, Mohammad W, Shah SM, Nawaz H, et al. (2005) Effect of tillage and fertilizer levels on wheat yield, nitrogen uptake and their correlation with carbon isotope discrimination under rainfed conditions in north-west Pakistan. *Soil and Tillage Research* 80: 47–57.
123. Shangquan ZP, Shao MA, Dyckmans J (2000) Nitrogen nutrition and water stress effects on leaf photosynthetic gas exchange and water use efficiency in winter wheat. *Environmental and Experimental Botany* 44: 141–149.
124. Clay DE, Engel RE, Long DS, Liu Z (2001) Nitrogen and Water Stress Interact to Influence Carbon-13 Discrimination in Wheat. *Soil Science Society of America Journal* 65: 1823–1828.
125. Winter CK, Davis SF (2006) Organic Foods. *Journal of Food Science* 71: R117–R124.
126. Schmidt H-I, Roßmann A, Voerkelius S, Schnitzler WH, Georgi M, et al. (2005) Isotope characteristics of vegetables and wheat from conventional and organic production. *Isotopes in Environmental and Health Studies* 41: 223–228.
127. Šturm M, Lojen S (2011) Nitrogen isotopic signature of vegetables from the Slovenian market and its suitability as an indicator of organic production. *Isotopes in Environmental and Health Studies* 47: 214–220.
128. Bateman AS, Kelly SD, Jickells TD (2005) Nitrogen and Sulfur Isotope Relationships between Crops and Fertilizer: Implications for Using Nitrogen Isotope Analysis as an Indicator of Agricultural Regime. *Journal of Agricultural and Food Chemistry* 53: 5760–5765.
129. Bateman AS, Kelly SD, Woolfe M (2007) Nitrogen Isotope Composition of Organically and Conventionally Grown Crops. *Journal of Agricultural and Food Chemistry* 55: 2664–2670.
130. Flores P, Fenoll J, Hellin P (2007) The Feasibility of Using  $\delta^{15}\text{N}$  and  $\delta^{13}\text{C}$  Values for Discriminating between Conventionally and Organically Fertilized Pepper (*Capsicum annuum* L.). *Journal of Agricultural and Food Chemistry* 55: 5740–5745.
131. Rapisarda P, Calabretta ML, Romano G, Intrigliolo F (2005) Nitrogen Metabolism Components as a Tool To Discriminate between Organic and Conventional Citrus Fruits. *Journal of Agricultural and Food Chemistry* 53: 2664–2669.
132. Rapisarda P, Camin F, Fabroni S, Perini M, Torrisi B, et al. (2010) Influence of Different Organic Fertilizers on Quality Parameters and the  $\delta^{15}\text{N}$ ,  $\delta^{13}\text{C}$ ,  $\delta^2\text{H}$ ,  $\delta^{34}\text{S}$ , and  $\delta^{18}\text{O}$  Values of Orange Fruit (*Citrus sinensis* L. Osbeck). *Journal of Agricultural and Food Chemistry* 58: 3502–3506.
133. Rogers KM (2008) Nitrogen Isotopes as a Screening Tool To Determine the Growing Regimen of Some Organic and Nonorganic Supermarket Produce from New Zealand. *Journal of Agricultural and Food Chemistry* 56: 4078–4083.
134. Camin F, Moschella A, Miselli F, Parisi B, Versini G, et al. (2007) Evaluation of markers for the traceability of potato tubers grown in an organic versus conventional regime. *Journal of the Science of Food and Agriculture* 87: 1330–1336.
135. Šturm M, Kacjan-Maršič N, Lojen S (2011) Can  $\delta^{15}\text{N}$  in lettuce tissues reveal the use of synthetic nitrogen fertiliser in organic production? *Journal of the Science of Food and Agriculture* 91: 262–267.
136. Flores P, Murray PJ, Hellin P, Fenoll J (2011) Influence of N doses and form on  $^{15}\text{N}$  natural abundance of pepper plants: considerations for using  $\delta^{15}\text{N}$  values

- as indicator of N source. *Journal of the Science of Food and Agriculture* 91: 2255–2258.
137. Schwarz HP (1991) Some theoretical aspects of isotope paleodiet studies. *Journal of Archaeological Science* 18: 261–275.
  138. Bogaard A, Heaton THE, Poulton P, Merbach I (2007) The impact of manuring on nitrogen isotope ratios in cereals: archaeological implications for reconstruction of diet and crop management practices. *Journal of Archaeological Science* 34: 335–343.
  139. Commisso RG, Nelson DE (2007) Patterns of plant  $\delta^{15}\text{N}$  values on a Greenland Norse farm. *Journal of Archaeological Science* 34: 440–450.
  140. Fraser RA, Bogaard A, Heaton T, Charles M, Jones G, et al. (2011) Manuring and stable nitrogen isotope ratios in cereals and pulses: towards a new archaeobotanical approach to the inference of land use and dietary practices. *Journal of Archaeological Science* 38: 2790–2804.
  141. Chepstow-Lusty AJ (2011) Agro-pastoral and social change in the Cuzco heartland of Peru: a brief history using environmental proxies. *Antiquity* 85: 570–582.
  142. Kubler G (1948) Towards Absolute Time: Guano Archaeology. *Memoirs of the Society for American Archaeology* 4: 29–50.
  143. Netherly PJ (1977) Local Level Lords on the North Coast of Peru [Unpublished Ph.D. Dissertation]. Ithaca: Cornell University. 366 p.
  144. Slovak NM, Paytan A (2011) Fisherfolk and farmers: Carbon and nitrogen isotope evidence from Middle Horizon Ancón, Peru. *International Journal of Osteoarchaeology* 21: 253–267.
  145. Knudson KJ, Aufderheide AE, Buikstra JE (2007) Seasonality and paleodiet in the Chiribaya polity of southern Peru. *Journal of Archaeological Science* 34: 451–462.
  146. Tomczak PD (2003) Prehistoric diet and socioeconomic relationships within the Osmore Valley of southern Peru. *Journal of Anthropological Archaeology* 22: 262–278.
  147. DeNiro MJ (1988) Marine food sources for prehistoric coastal Peruvian camelids: isotopic evidence and implications. In: Wing ES, Wheeler JC, eds. *Economic Prehistory of the Central Andes*. Oxford: British Archaeological Reports International Series 427. pp 119–128.
  148. Cushman GT The Lords of Guano: Science and the Management of Peru's Marine Environment, 1800–1973 [Unpublished Ph.D. Dissertation]: University of Texas, Austin. 762 p.
  149. Cordle C (2007) The Guano Voyages. *Rural History* 18: 119–133.
  150. Simmons D (2006) Waste Not, Want Not: Excrement and Economy in Nineteenth-Century France. *Representations* 96: 73–98.
  151. Cox G, Sealy J (1997) Investigating Identity and Life Histories: Isotopic Analysis and Historical Documentation of Slave Skeletons Found on the Cape Town Foreshore, South Africa. *International Journal of Historical Archaeology* 1: 207–224.
  152. Katzenberg MA (1995) Nitrogen isotope evidence for weaning age in a nineteenth century Canadian skeletal sample. In: Grauer AL, ed. *Bodies of Evidence: Reconstructing History through Skeletal Analysis*. Cambridge: Wiley-Liss. pp 221–235.
  153. Cox G, Sealy J, Schrire C, Morris A (2001) Stable Carbon and Nitrogen Isotopic Analyses of the Underclass at the Colonial Cape of Good Hope in the Eighteenth and Nineteenth Centuries. *World Archaeology* 33: 73–97.
  154. Klippel WE (2001) Sugar Monoculture, Bovid Skeletal Part Frequencies, and Stable Carbon Isotopes: Interpreting Enslaved African Diet at Brimstone Hill, St Kitts, West Indies. *Journal of Archaeological Science* 28: 1191–1198.
  155. Valentin F, Bocherens H, Gratuz B, Sand C (2006) Dietary patterns during the late prehistoric/historic period in Cikobia island (Fiji): insights from stable isotopes and dental pathologies. *Journal of Archaeological Science* 33: 1396–1410.
  156. Katzenberg M, Saunders S, Abonyi S (2000) Bone Chemistry, Food and History: A Case Study from 19th Century Upper Canada. In: Ambrose SH, Katzenberg MA, eds. *Biogeochemical Approaches to Paleodietary Analysis*. New York: Kluwer Academic. pp 1–22.
  157. Roy DM, Hall R, Mix AC, Bonnichen R (2005) Using stable isotope analysis to obtain dietary profiles from old hair: A case study from Plains Indians. *American Journal of Physical Anthropology* 128: 444–452.
  158. Schroeder H, O'Connell TC, Evans JA, Shuler KA, Hedges REM (2009) Trans-Atlantic slavery: Isotopic evidence for forced migration to Barbados. *American Journal of Physical Anthropology* 139: 547–557.
  159. Sealy J, Armstrong R, Schrire C (1995) Beyond lifetime averages: tracing life histories through isotopic analysis of different calcified tissues from archaeological human skeletons. *Antiquity* 69: 290–300.
  160. Hobson KA, Clark RG (1992) Assessing avian diets using stable isotopes II: factors influencing diet-tissue fractionation. *The Condor* 94: 189–197.
  161. Hobson KA (1995) Reconstructing Avian Diets Using Stable-Carbon and Nitrogen Isotope Analysis of Egg Components: Patterns of Isotopic Fractionation and Turnover. *The Condor* 97: 752–762.
  162. Wainright SC, Haney JC, Kerr C, Golovkin AN, Flint MV (1998) Utilization of nitrogen derived from seabird guano by terrestrial and marine plants at St. Paul, Pribilof Islands, Bering Sea, Alaska. *Marine Biology* 131: 63–71.
  163. Mizutani H, Kabaya Y, Wada E (1985) High-performance liquid chromatographic isolation of uric acid from soil for isotopic determination. *Journal of Chromatography A* 331: 371–381.
  164. Bird MI, Tait E, Wurster CM, Furness RW (2008) Stable carbon and nitrogen isotope analysis of avian uric acid. *Rapid Communications in Mass Spectrometry* 22: 3393–3400.
  165. Kirshenbaum I, Smith JS, Crowell T, Graff J, McKee R (1947) Separation of the Nitrogen Isotopes by the Exchange Reaction between Ammonia and Solutions of Ammonium Nitrate. *Journal of Chemical Physics* 15: 440–446.
  166. Shearer G, Kohl DH (1986)  $\text{N}_2$ -Fixation in Field Settings: Estimations Based on Natural  $^{15}\text{N}$  Abundance. *Australian Journal of Plant Physiology* 13: 699–756.
  167. Feigin A, Shearer G, Kohl DH, Commoner B (1974) The Amount and Nitrogen-15 Content of Nitrate in Soil Profiles from two Central Illinois Fields in a Corn-Soybean Rotation. *Soil Science Society of America Journal* 38: 465–471.
  168. Hoch MP, Fogel ML, Kirchman DL (1992) Isotope Fractionation Associated with Ammonium Uptake by a Marine Bacterium. *Limnology and Oceanography* 37: 1447–1459.
  169. Yoneyama T, Fujiwara H, Wilson JW (1998) Variations in fractionation of carbon and nitrogen isotopes in higher plants: N-metabolism and partitioning in phloem and xylem. In: Griffiths H, ed. *Stables Isotopes, Integration of Biological, Ecological and Geochemical Processes*. Oxford: BIOS Scientific Publishers. pp 99–109.
  170. Mariotti A, Mariotti F, Champigny M-L, Amarger N, Moyse A (1982) Nitrogen Isotope Fractionation Associated with Nitrate Reductase Activity and Uptake of  $\text{NO}_3^-$  by Pearl Millet. *Plant Physiology* 69: 880–884.
  171. Miller A, Cramer M (2005) Root Nitrogen Acquisition and Assimilation. *Plant and Soil* 274: 1–36.
  172. Tcherkez G, Farquhar GD (2006) Isotopic fractionation by plant nitrate reductase, twenty years later. *Functional Plant Biology* 33: 531–537.
  173. Barrett K, Anderson WB, Wait DA, Grismer LL, Polis GA, et al. (2005) Marine subsidies alter the diet and abundance of insular and coastal lizard populations. *Oikos* 109: 145–153.
  174. Bokhorst S, Huiskes A, Convey P, Aerts R (2007) External nutrient inputs into terrestrial ecosystems of the Falkland Islands and the Maritime Antarctic region. *Polar Biology* 30: 1315–1321.
  175. Cocks MP, Balfour DA, Stock WD (1998) On the uptake of ornithogenic products by plants on the inland mountains of Dronning Maud Land, Antarctica, using stable isotopes. *Polar Biology* 20: 107–111.
  176. Erskine PD, Bergstrom DM, Schmidt S, Stewart GR, Tweedie CE, et al. (1998) Subantarctic Macquarie Island – a model ecosystem for studying animal-derived nitrogen sources using  $^{15}\text{N}$  natural abundance. *Oecologia* 117: 187–193.
  177. Harding JS, Hawke DJ, Holdaway RN, Winterbourn MJ (2004) Incorporation of marine-derived nutrients from petrel breeding colonies into stream food webs. *Freshwater Biology* 49: 576–586.
  178. Hawke DJ, Holdaway RN (2005) Avian assimilation and dispersal of carbon and nitrogen brought ashore by breeding Westland petrels (*Procellaria westlandica*): a stable isotope study. *Journal of Zoology* 266: 419–426.
  179. Hawke DJ, Newman J (2007) Carbon-13 and nitrogen-15 enrichment in coastal forest foliage from nutrient-poor and seabird-enriched sites in southern New Zealand. *New Zealand Journal of Botany* 45: 309–315.
  180. Hobara S, Koba K, Osono T, Tokuchi N, Ishida A, et al. (2005) Nitrogen and phosphorus enrichment and balance in forests colonized by cormorants: Implications of the influence of soil adsorption. *Plant and Soil* 268: 89–101.
  181. Kameda K, Koba K, Hobara S, Osono T, Terai M (2006) Pattern of natural  $^{15}\text{N}$  abundance in lakeside forest ecosystem affected by cormorant-derived nitrogen. *Hydrobiologia* 567: 69–86.
  182. Kolb G, Jerling L, Hambäck P (2010) The Impact of Cormorants on Plant-Arthropod Food Webs on Their Nesting Islands. *Ecosystems* 13: 353–366.
  183. Markwell TJ, Daugherty CH (2002) Invertebrate and lizard abundance is greater on seabird-inhabited islands than on seabird-free islands in the Marlborough Sounds, New Zealand. *Ecoscience* 9: 293–299.
  184. Mizota C, Naikatin A (2007) Nitrogen isotope composition of inorganic soil nitrogen and associated vegetation under a sea bird colony on the Hatana islands, Rotuma Group, Fiji. *Geochemical Journal* 41: 297–301.
  185. Mizota C (2009) Nitrogen isotopic patterns of vegetation as affected by breeding activity of Black-tailed Gull (*Larus crassirostris*): A coupled analysis of feces, inorganic soil nitrogen and flora. *Applied Geochemistry* 24: 2027–2033.
  186. Mizutani H, Wada E (1988) Nitrogen and Carbon Isotope Ratios in Seabird Rookeries and their Ecological Implications. *Ecology* 69: 340–349.
  187. Mizutani H, Kabaya Y, Moors PJ, Speir TW, Lyon GL (1991) Nitrogen Isotope Ratios Identify Deserted Seabird Colonies. *The Auk* 108: 960–964.
  188. Stapp P, Polis GA, Sanchez Pinero F (1999) Stable isotopes reveal strong marine and El Niño effects on island food webs. *Nature* 401: 467–469.
  189. Young HS, McCauley DJ, Dunbar RB, Dirzo R (2010) Plants cause ecosystem nutrient depletion via the interruption of bird-derived spatial subsidies. *Proceedings of the National Academy of Sciences* 107: 2072–2077.
  190. Zhu R, Liu Y, Ma E, Sun J, Xu H, et al. (2009) Nutrient compositions and potential greenhouse gas production in penguin guano, ornithogenic soils and seal colony soils in coastal Antarctica. *Antarctic Science* 21: 427–438.
  191. Bateman AS, Kelly SD (2007) Fertilizer nitrogen isotope signatures. *Isotopes in Environmental and Health Studies* 43: 237–247.
  192. Dijkstra P, Menyailo OV, Doucet RR, Hart SC, Schwartz E, et al. (2006) C and N availability affects the  $^{15}\text{N}$  natural abundance of the soil microbial

- biomass across a cattle manure gradient. *European Journal of Soil Science* 57: 468–475.
193. Kerley SJ, Jarvis SC (1996) Preliminary studies of the impact of excreted N on cycling and uptake of N in pasture systems using natural abundance stable isotopic discrimination. *Plant and Soil* 178: 287–294.
  194. Lim S-S, Choi W-J, Kwak J-H, Jung J-W, Chang S, et al. (2007) Nitrogen and carbon isotope responses of Chinese cabbage and chrysanthemum to the application of liquid pig manure. *Plant and Soil* 295: 67–77.
  195. Yun S-I, Ro H-M, Choi W-J, Chang SX (2006) Interactive effects of N fertilizer source and timing of fertilization leave specific N isotopic signatures in Chinese cabbage and soil. *Soil Biology and Biochemistry* 38: 1682–1689.
  196. Choi W-J, Ro H-M, Hobbie EA (2003) Patterns of natural  $^{15}\text{N}$  in soils and plants from chemically and organically fertilized uplands. *Soil Biology and Biochemistry* 35: 1493–1500.
  197. Choi W-J, Ro H-M (2003) Differences in isotopic fractionation of nitrogen in water-saturated and unsaturated soils. *Soil Biology and Biochemistry* 35: 483–486.
  198. del Amor FM, Navarro J, Aparicio PM (2008) Isotopic Discrimination as a Tool for Organic Farming Certification in Sweet Pepper. *Journal of Environmental Quality* 37: 182–185.
  199. Nakano A, Uehara Y (2007) Effects of different kinds of fertilizer and application methods on  $\delta^{15}\text{N}$  values of tomato. *Japan Agricultural Research Quarterly* 41: 219–226.